



This document is a postprint version of an article published in Harmful Algae© Elsevier after peer review. To access the final edited and published work see <https://doi.org/10.1016/j.hal.2020.101913>

Document downloaded from:



1 ***Gambierdiscus* and *Fukuyoa* as potential indicators of ciguatera risk in the Balearic Islands**

2 **Authors:** Àngels Tudó^{a,b}, Anna Toldrà^a, Maria Rey^a, Irene Todolí^a, Karl B. Andree^a, Margarita
3 Fernández-Tejedor^a, Mònica Campàs^a, Francesc X. Sureda^b, Jorge Diogène^a

4 ^aIRTA, Ctra. Poble Nou Km 5.5, 43540, Sant Carles de la Ràpita, Tarragona, Spain.

5 ^bPharmacology Unit, Faculty of Medicine and Health Sciences, Universitat Rovira i Virgili, C/St. Llorenç
6 21, E-43201, Reus (Tarragona), Spain.

7 **Corresponding author:** jorge.diogene@irta.cat

8 **Highlights**

- 9 • *G. australes* and *F. paulensis* are well distributed and established in the Balearic Islands, a
10 region free of Ciguatera Poisoning.
- 11 • Overall, low CTX-like toxicity was detected in *G. australes* and *F. paulensis* strains.
- 12 • Presence of MTX-like activity was detected in *G. australes* strains.

13 **Keywords**

14 Ciguatera, ciguatoxins, maitotoxins, *Gambierdiscus*, *Fukuyoa*, neuro-2a cell-based assay.

15 **Abstract**

16 *Gambierdiscus* and *Fukuyoa* are genera of toxic dinoflagellates which were mainly considered as
17 endemic to marine intertropical areas, and that are well known as producers of ciguatoxins (CTXs)
18 and maitotoxins (MTXs). Ciguatera poisoning (CP) is a human poisoning occurring after the
19 consumption of fish or more rarely, shellfish containing CTXs. The presence of these microalgae in a
20 coastal area is an indication of potential risk of CP. This study assesses the risk of CP in the Balearic
21 Islands (Western Mediterranean Sea) according to the distribution of both microalgae genera, and
22 the presence of CTX-like and MTX-like toxicity in microalgal cultures as determined by neuro-2a cell

23 based-assay (neuro-2a CBA). Genetic identification of forty-three cultured microalgal strains isolated
24 from 2016 to 2019 revealed that all of them belong to the species *G. australes* and *F. paulensis*. Both
25 species were widely distributed in Formentera, Majorca and Minorca. Additionally, all strains of *G.*
26 *australes* and two of *F. paulensis* exhibited signals of CTX-like toxicity ranging respectively between
27 1-380 and 8-16 fg CTX1B equivalents (equiv.) · cell⁻¹. Four extracts of *F. paulensis* exhibited a novel
28 toxicity response in neuro-2a cells consisting of the recovery of the cell viability in the presence of
29 ouabain and veratridine. In addition, *G. australes* showed MTX-like toxicity while *F. paulensis* strains
30 did not. Overall, the low CTX-like toxicities detected indicate that the potential risk of CP in the
31 Balearic Islands is low, although, the presence of CTX-like and MTX-like toxicity in those strains reveal
32 the necessity to monitor these genera in the Mediterranean Sea.

33 **1. Introduction**

34 *Gambierdiscus* (Adachi and Fukuyo, 1979) and *Fukuyoa* (Gómez et al., 2015) (Dinophyceae) are
35 marine benthic dinoflagellates that live attached to different substrates such as macroalgae, corals,
36 rocks and sands in well-illuminated habitats but also at very low light levels (>45 m depth) (Tester et
37 al., 2013). Historically, the genera *Gambierdiscus* and *Fukuyoa* were known to be distributed
38 primarily in tropical and subtropical areas of the Caribbean Sea, the Pacific and Indian Ocean.
39 However, in recent decades, both genera have been reported in warm-temperate areas. The genus
40 *Gambierdiscus* was recently recorded in the North East Atlantic Ocean (Fernández-Zabala et al., 2019;
41 Fraga et al., 2011; Rodríguez et al., 2017), North West Atlantic Ocean (Litaker et al., 2009), South
42 West Atlantic (Nascimento et al., 2015), the Mediterranean Sea (Aligizaki and Nikolaidis, 2008; Tudó
43 et al., 2018), the Red Sea (Catania et al., 2017), Sea of Japan (Jang et al., 2018) and the South Pacific
44 Ocean (Kohli et al., 2014a; Larsson et al., 2018). In contrast species of the genus *Fukuyoa* (formerly
45 within the genus *Gambierdiscus*), have been reported in the Atlantic Ocean (Gómez et al., 2015), the
46 Mediterranean Sea (Laza-Martínez et al., 2016; Aligizaki et al., 2018), the South Pacific Ocean (Rhodes
47 et al., 2017), the China Sea and the Asia Pacific region (Larsson et al., 2019, 2018; Leung et al., 2018).

48 *Gambierdiscus* and *Fukuyoa* produce multiple secondary metabolites, among which are included
49 ciguatoxins (CTXs) and maitotoxins (MTXs) (Chinain et al., 2010; Holmes et al., 1990; Lewis and
50 Holmes, 1993; Munday et al., 2017; Satake et al., 1996). CTXs are lipophilic polyethers, that bind to
51 voltage-gated sodium channels (VGSCs), thereby inhibiting the inactivation process of VGSCs
52 resulting in intracellular sodium increase (Hidalgo et al., 2002; Molgó et al., 1993; Nicholson and
53 Lewis, 2006; Strachan et al., 1999). Moreover, CTXs are potassium channel inhibitors (Inserra et al.,
54 2017). MTXs are amphiphilic polyethers that bind to Ca²⁺ independent voltage gated channels and
55 non-selective ion channels causing an increase of intracellular Ca²⁺ (Reyes et al., 2014).

56 CTXs in fish or shellfish are responsible for the human intoxication known as Ciguatera Poisoning (CP)
57 (Bagnis, 1993; Bagnis et al., 1980). CTXs enter marine food webs through invertebrates and
58 herbivorous fish, where they may be biotransformed along the food webs and bioaccumulated at
59 different trophic levels, eventually reaching humans (Bagnis et al., 1980; Yasumoto et al., 1977).

60 Regarding MTXs, their implication in CP is unlikely. Although its intraperitoneal administration in mice
61 is more toxic than CTXs, their oral potency is almost non-detectable (Munday et al., 2017). In
62 addition, their bioaccumulation along the food webs is low (Litaker et al., 2010; Munday, 2014;
63 Yasumoto et al., 1971) and they have not been found in the tissue of fish involved in CP cases.

64 However, snapper (*Chrysophrys auratus*) (previously *Pagrus auratus*), that had been experimentally
65 fed with *G. australes* contained MTXs in their viscera, liver and muscle (Kohli et al., 2014b).

66 Although, epidemiological records of CP are not available at a global level, it is estimated that CP
67 affects between 25,000 – 500,000 people per year (Fleming et al., 1998; Friedman et al., 2017;
68 Skinner et al., 2011). CP effects include gastrointestinal, neurological, and cardiovascular symptoms,
69 and the latter two can last for months or years (Friedman et al., 2017). Fatal cases of CP are rare
70 (Chan, 2016; Diogène et al., 2017). CP occurs mainly in tropical and subtropical areas (35 °N - 35 °S),
71 but in more recent decades, CP cases have been reported in temperate areas, previously free of CP
72 (Bravo et al., 2015; Chinain et al., 2019; Gouveia et al., 2010).

73 In the Mediterranean Sea, the presence of CTXs in fish, or confirmed CP cases have not been
74 demonstrated. Follow-up investigations of previous descriptions of CP cases in the eastern
75 Mediterranean did not find CTXs in fish tissue (Bentur and Spanier, 2007; Herzberg, 1973; Raikhlin-
76 Eisenkraft and Bentur, 2002; Raikhlin-Eisenkraft et al., 1988; Spanier et al., 1989). The detection of
77 possible CTX-compounds in *Siganus* sp. by Bentur and Spanier (2007) was performed using a *Cigua-*
78 *Check* strip test, which was later considered unreliable (Bienfang et al., 2011). In addition, the clinical
79 symptoms described, including hallucinations, are rare in CP cases (Chinain et al., 2019) and they are
80 indicative of ichthyallyeinotoxism, which is often mistaken for cases of CP (De Haro and Pommier,
81 2006).

82 At present, five confirmed species of the genus *Gambierdiscus* and *Fukuyoa* live in the Mediterranean
83 Sea (Aligizaki et al., 2018; Laza-Martínez et al., 2016; Litaker et al., 2009; Tudó et al., 2018). The
84 presence of certain CTX-producing species in the area can be indicative of a higher risk of CP in
85 comparison to areas where they are absent (Chinain et al., 2019, 2010; Friedman et al., 2017).
86 Nonetheless, evaluating CTX-production by these species is important to estimate the risk, since CTX
87 production varies according to species, and high and low CTX-producers species have been
88 characterized (Litaker et al., 2017; Pisapia et al., 2017). For the estimation of CTX production in
89 *Gambierdiscus*, growth phases and strain variability among isolates of the same species have to be
90 taken into account (Reverté et al., 2018; Rossignoli et al., 2020).

91 The goal of this study was to assess the potential risk of CP based on the presence in the Balearic
92 Islands of the genera *Gambierdiscus* and *Fukuyoa* (Western Mediterranean Sea), and their potential
93 production of compounds with CTX-like and MTX-like activity. This is the first study that provides
94 information about the risk of CP in the Balearic Islands, according to the presence of the genera
95 *Gambierdiscus* and *Fukuyoa* in several sampling locations, and their evaluation of toxin production
96 of several strains.

97 **2. Materials and Methods**

98 **2.1 Reagents and equipment**

99 CTX1B was provided by Dr. Lewis, University of Queensland (Lewis et al., 1991). Neuroblastoma
100 murine cells (neuro-2a) were purchased from ATCC LGC standards (USA). Poly-L-lysine, foetal bovine
101 serum (FBS), L-glutamine solution, ouabain, veratridine, phosphate buffered saline (PBS), penicillin,
102 streptomycin, RPMI-1640 medium, sodium pyruvate, thiazolyl blue tetrazolium bromide (MTT) and
103 SKF96365 were purchased from Merck KGaA (Germany). Dimethyl sulfoxide (DMSO) and absolute
104 methanol were purchased from Honeywell (Spain) and Chemlab (Spain) respectively. Taq Polymerase
105 was purchased from Invitrogen (Spain). QIAquick PCR Purification Kit was obtained from Qiagen
106 (Germany).

107 **2.2 Sampling, cell isolation and initial culturing**

108 The Balearic Archipelago (North West Mediterranean Sea) is located at 170 km distance from the
109 Iberian Peninsula (Fig. 1). It is characterized by a narrow continental shelf surrounding a rocky coast,
110 with occasional sea grass meadows over a biogenic muddy bottom. Samples from the Balearic Islands
111 were collected at different islands, specifically, in Formentera in late September 2016, in Majorca
112 and Minorca in early September 2017 and in early October 2018. In Minorca an additional sampling
113 was performed in late September 2019. At each sampling point, two different types of samples were
114 collected: 1) epilithic, which were obtained by scraping of the substrate (rocks) with a plastic bottle
115 (Nalgene, HDPE, 1L), and 2) epiphytic, which were obtained from macroalgae that were collected
116 using plastic bottles under water. Macroalgae were identified morphologically at the genus level.
117 Each sample was kept in the container and was intensively shaken by hand to release the
118 dinoflagellates from the substrates. Samples were sieved through a 200 μm nylon mesh. The filtered
119 water was stored in two plastic bottles (Nalgene, HDPE, 125mL), one with 125 mL was kept untreated
120 to isolate live cells and another was preserved in 3% Lugol's iodine solution for further observation
121 in the laboratory. Coordinates of each sampling station were recorded by GPS. Salinity, oxygen (%
122 and $\text{mg} \cdot \text{L}^{-1}$), temperature and pH were recorded in situ using a multiparametric probe (YSI 556 MPS).

123 Samples were observed under an inverted light microscope Leica DMIL (Leica Microsystems GmbH,
124 Germany) and individual microalgal cells were isolated by capillary method (Hoshaw and Rosowski,
125 1973) to establish clonal cultures. Each cell was inoculated in a well of an untreated Nunc 24 well
126 plate (Thermo Fisher Scientific) with 1 mL of modified ES medium (Provasoli, 1968). Medium was
127 prepared from sterile aged seawater from L'Ametlla de Mar (Spain), Mediterranean Sea (40.8465° N;
128 0.77243° E) and salinity was adjusted to 36. After 2-3 weeks, when cell abundance of cultures reached
129 20-30 cells · mL⁻¹, cells were transferred to 28 mL round bottom glass tubes (Thermo Fisher Scientific)
130 containing 10 mL of medium. Cultures were maintained in a culture chamber at a temperature of 24
131 ± 0.5 °C, which is the average of the range of the optimal temperatures of growth for *G. australes*
132 (Yoshimatsu et al., 2014) and in coherence with our previous studies, Reverté et al. (2018), Caillaud
133 et al. (2010). Illumination in a 12:12 light:dark cycle was provided by fluorescent tubes with white
134 light and with photon irradiance of 100 μmol photons · m⁻² · s⁻¹ measured by an irradiator (QSL-
135 2100 Radiometer, Biospherical Instruments, San Diego, USA). Preserved field samples were settled
136 in 10 mL sedimentation chambers and observed under an inverted light microscope for microalgal
137 identification.

138 **2.3 Molecular identification**

139 Molecular identification at species level was performed by sequencing the D8-D10 region of the 28S
140 ribosomal large subunit gene (LSU rDNA). Molecular identification was conducted for 34
141 *Gambierdiscus* strains and 9 *Fukuyoa* strains. To that purpose, strains were inoculated in 50mL of
142 medium at 50 cells · mL⁻¹ in 25 cm² sterile Nunclon™ culture flasks (Thermo Fisher Scientific), and
143 when cultures achieved the exponential phase, they were harvested by centrifugation at 4300g for
144 20 min (Allegra X-15R, Beckman Coulter). Genomic DNA was extracted by
145 phenol/chloroform/isoamylalcohol (PCI) extraction following Toldrà et al., (2018). After DNA
146 extraction, genomic DNA was quantified and checked for its purity using a NanoDrop 2000
147 spectrophotometer (Thermo Fisher Scientific) and stored at -20 °C. Afterwards, the region D8-D10
148 was amplified by PCR using the primers FD8 and RB (Chinain et al., 1999). Each 25 μL reaction mixture

149 contained 600 μM dNTP, 2 mM MgCl_2 , 0.2 μM of each primer, 1 U of Taq polymerase, 5% DMSO, and
150 0.4–2 $\text{ng} \cdot \text{ul}^{-1}$ of DNA template. Amplifications were carried out in a Mastercycler nexus gradient
151 thermal cycler (Eppendorf, Spain) as follows: an initial denaturation step of 5 min at 95 °C, 40 cycles
152 of 30 s at 95 °C, 45 s at 60 °C, and 30 s at 72 °C and a final extension step of 10 min at 72 °C. Each PCR
153 reaction was verified by agarose gel electrophoresis and visualized with ethidium bromide stain. The
154 resulting PCR products of ~ 840–910 bp were purified with the QIAquick PCR Purification Kit. Purified
155 products were bi-directionally sequenced by an external company (Sistemas Genómicos, LLC,
156 Valencia, Spain). Consensus sequences obtained from both reads for each strain were manually
157 edited using BioEdit v7.0.5.2 (Hall, 1999) and deposited in GenBank. Sequences were aligned using
158 MAFFT v.7 (Rozewicki et al., 2019) with G-INS-1 progressive method. The final alignment consisted
159 of 617 positions. The evolutionary model of data was estimated using jModelTest 2.1.10 (Darriba et
160 al., 2012) and the phylogenetic relationships were inferred by Maximum likelihood (ML) using RaxML
161 v.8 (Stamatakis, 2014) and Bayesian inference (BI) using Mr. Bayes v.3.2.2 (Huelsenbeck and
162 Ronquist, 2001). In the BI approach two analyses were run in parallel, 10^6 generations, and four
163 chains in each run. The parameters used for analysis were nst=mixed and rates=gamma. By default,
164 25% of the trees were discarded. Stability of the chains were checked using Tracer v.1.7.1 (Rambaut
165 et al., 2018).

166 **2.4 Morphological characterization**

167 For morphological characterization strains were acclimated at least one year to avoid stress-induced
168 variance during the adaptation period to laboratory conditions (Bomber et al., 1989).

169 **2.4.1 Light microscopy (LM)**

170 Seven monoclonal cultures of *G. australes* and two of *F. paulensis* were inoculated at 20–30 cells $\cdot \text{mL}^{-1}$
171 ¹ in 28 mL round bottom glass tubes. When cultures arrived at final exponential phase (after \pm 20
172 days) cells were stained with Calcofluor White M2R (Sigma Aldrich, Spain) according to Fritz and
173 Triemer (1985). Calcofluor-stained cells were observed using an epifluorescence microscope (LEICA

174 DMLB and NIKON eclipse 80i) equipped with an Olympus camera (Olympus DP70), and they were
175 measured using the software Olympus DP controller (Olympus Corporation). Morphological
176 characteristics of microalgal cells were based on the tabulation system described in Fraga et al.
177 (2011). Cell length was determined as the apical to antapical distance dimensions, depth as the
178 dorso-ventral distance and width as the transdiameter distance that is the longest distance between
179 opposed sides of the cingulum (Balech, 1989). Cell dimensions were expressed as mean \pm standard
180 deviation (SD).

181 **2.4.2 Scanning electron microscopy (SEM)**

182 SEM was used to study two monoclonal cultures of *G. australes* (IRTA-SMM-17-253 and IRTA-SMM-
183 17-164) and one of *F. paulensis* (IRTA-SMM-17-211). For that, ten mL samples of cultures at the initial
184 exponential growth phase were fixed with glutaraldehyde at a final concentration of 4% during 2 h
185 at room temperature. After that, 3 mL of culture were collected with a syringe by applying a low
186 pressure on 5 μ m Nuclepore Track-Etch Membrane (Thermo Fisher Scientific) coated by poly-L-lysine
187 and held in a plastic filter mould 13 mm (PALL, life Science). Filters were rinsed twice. Once with
188 seawater (autoclaved and filtered by active carbon 0.2 μ m) and a second time with filtered
189 seawater/MilliQ water (50:50, v:v). Afterwards, filters were rinsed in a graded EtOH series of 30, 50,
190 70, 80, 90 and twice with 96% (v:v). Later, filters were kept in a recipient with absolute EtOH and they
191 were sent to the Scanning Electron Microscopy Service in the Institute of Marine Science (ICM-CSIC).
192 In the facilities, filters were submitted to critical-point drying with liquid carbon dioxide in a BAL-TEC
193 CPD030 unit (Leica Microsystems, Austria). Dried filters were mounted on stubs with colloidal silver,
194 then sputter-coated with gold in a Q150R S (Quorum Technologies Ltd). Cells were observed with a
195 Hitachi S3500N scanning electron microscope (Hitachi High Technologies Co., Ltd, Japan) at an
196 accelerating voltage of 5 kV. Length and width of the Po plate and the second antapical plate, 2''
197 plate (Fraga et al. 2011), were measured and the number of pores of the Po plate were counted.
198 Measurements were made using ImageJ software (Schneider et al., 2012).

199 **2.5 Growth dynamics analysis**

200 Before the growth dynamics analysis strains were first acclimated to laboratory conditions for
201 approximately 1 year. To evaluate growth dynamics, three strains of *G. australes* (IRTA-SMM-17-162,
202 IRTA-SMM-17-189, IRTA-SMM-17-271) and one strain of *F. paulensis* (IRTA-SMM-17-209) were
203 randomly selected from the algal collection. For each strain, 500 mL of medium were inoculated into
204 1.5 L Fernbach flasks at an initial concentration of $50 \text{ cells} \cdot \text{mL}^{-1}$ in triplicate. Every 2-3 days, at the
205 same time of the day, each culture was vigorously manually homogenized and 3 mL samples from
206 each replicate were collected and preserved with 3% Lugol's iodine solution. Three countings of each
207 sample were conducted under observation in an inverted light microscope using a 0.5 mL Kolkwitz
208 counting chamber (Plankton Chamber acc. to Kolkwitz-Hydro-bios). For each day and replicate,
209 average of the cell abundance ($\text{cells} \cdot \text{mL}^{-1}$) and SD were estimated. The growth rate (r) of each
210 replicate was estimated by the equation of a linear regression by the least square fit after logarithmic
211 transformation of the cell abundance vs time considering at least 3 points of the exponential phase.
212 The growth rate was expressed in units of divisions ($\text{div.} \cdot \text{day}^{-1}$). Moreover, the doublings per day (K)
213 were calculated as $K = r / \ln(2)$ (Eq. 1) and expressed as doublings day^{-1} (Guillard, 1973). Besides, the
214 time of division or doubling time (T_d) was calculated as $T_d = \ln(2) / r$ (Eq. 2) and expressed as day^{-1}
215 (Guillard, 1973). Also, growth phases were defined as Wood et al. (2005) where the exponential
216 phase (log phase) was defined as the period when the slope of the regression line between elapsed
217 time and log cell concentration was maximum. Late-exponential (late log) – early stationary phase
218 was defined when the slope of the regression line between elapsed time and log cell concentration
219 is reduced in comparison to the slope from the log phase. The negative growth was defined by
220 constituent decrease of cells, which was assessed by observation of microalgal cells by light
221 microscope and confirmed by observation of empty thecae. However, cultures did not arrive at
222 significantly negative growth.

223 **2.6 Toxin Analysis**

224 **2.6.1 Culture, harvesting and algal extraction**

225 The CTX-like activity was evaluated for 21 strains of *G. australes* (11 strains from Majorca and 10 from
226 Minorca), and 6 strains of *F. paulensis* (2 strains from Majorca and 4 from Minorca) harvested at late
227 log - early stationary phases of culture. For this purpose, strains were inoculated in 500 mL of medium
228 in 1.5 L Fernbach flasks at an initial concentration of $50 \text{ cells} \cdot \text{mL}^{-1}$. When culture arrived at late-
229 exponential phase (after 20 ± 3 days), cultures were vigorously shaken, and 15 mL aliquots were fixed
230 using Lugol's iodine solution (3%) to estimate the cell concentration ($\text{cell} \cdot \text{mL}^{-1}$) in the culture.
231 Subsequently, the remaining volume was collected in sterile 50 mL Falcon tubes and centrifuged at
232 4300 g for 20 min. Supernatants were discarded, and pellets were pooled in one 50 mL Falcon tube.
233 Centrifugation was repeated, the supernatant was discarded, and the pellet was kept at $-20 \text{ }^\circ\text{C}$ with
234 absolute methanol (10 mL for 10^6 cells) until toxin extraction.

235 To prepare microalgal extracts, cell pellets of approximately 5×10^5 to 10^6 cells with methanol were
236 sonicated using an ultrasonic cell disrupter (Watt ultrasonic processor VCX750, USA). The tip
237 amplitude was set at 37% 3 sec on/2 sec off for 15 minutes. The sample was then centrifuged at 600
238 g for 5 min at $4 \text{ }^\circ\text{C}$. Supernatant was transferred to a glass vial. Procedure was repeated twice, one
239 with methanol and another with aqueous methanol (50:50; v:v) (10 mL for 10^6 cells). The methanol
240 extracts were then evaporated to dryness with a rotary evaporator (Büchi Syncore, Switzerland) or
241 dried under N_2 gas (Turbovap, Caliper, Hopkinton, USA) at $40 \text{ }^\circ\text{C}$. The aqueous methanol was
242 evaporated at $70 \text{ }^\circ\text{C}$. When dryness was achieved, absolute methanol was added to the glass vials,
243 then extracts were pooled, filtered with PTFE filters ($0.2 \mu\text{m}$) and stored at $-20 \text{ }^\circ\text{C}$.

244 **2.6.2 CTX-like toxicity evaluation**

245 The presence of CTX-like activity was evaluated on microalgal pellets of 21 cultures of *G. australes*,
246 and 6 cultures of *F. paulensis* harvested at late log - early stationary phase. The evaluation was
247 conducted using the neuro-2a CBA. This assay is used to detect bioactive compounds which target
248 the voltage gated sodium channels (VGSCs) (Cañete and Diogène, 2008; Manger et al., 2003, 1995,
249 1993). Ouabain blocks the sodium efflux through the inhibition of the $\text{Na}^+/\text{K}^+ \text{ -ATPase}$ pump

250 (Catterall and Nirenberg, 1973) whereas, the veratridine blocks the sodium voltage-gate channel in
251 an open position (Catterall, 1986). The cell viability of the neuro-2a cells is affected when the extract
252 contains CTXs or CTX-like compounds (molecules that activates to VGSC) after the ouabain and
253 veratridine treatment (Cañete and Diogène, 2008; Manger et al., 2003, 1995, 1993). Exposure of
254 neuro-2a cells to CTX1B standard (reference) or microalgal extracts was performed following the
255 protocol described in Reverté et al. (2018). Briefly, neuro-2a cells were seeded at a density of $1.4 \times$
256 10^5 cells \cdot mL⁻¹ in 96-well plates. After 24 h, ouabain and veratridine (O/V) were added to a final
257 concentration at 140 μ M and 14 μ M respectively, then, 10 μ L of each sample (serial dilutions of
258 extract or standard) was added to each well in triplicate. Concentrations of CTX1B ranged between
259 0.2 to 25 pg \cdot mL⁻¹ and concentrations of microalgal extracts ranged between 0.3 to 1000 cells equiv.
260 mL⁻¹ for *G. australes* and 10 to 4000 cells equiv. \cdot mL⁻¹ for *F. paulensis*. After 24 h, cell viability was
261 measured using a colorimetric assay using 3-(4,5-dimethylthiazol-2-yl)-2,5 diphenyltetrazolium
262 (Manger et al. 1993). Absorbance was measured at 570 nm using an automated plate
263 spectrophotometer (Synergy HT, Biotek, USA).

264 Hence, for every assay a calibration curve of cell viability with the standard was obtained. Curves
265 were adjusted to a sigmoidal logistic 4-parameter regression using SigmaPlot software 12.0 (Systat
266 Software Inc., USA). Limit of detection (LOD) was calculated as the necessary concentration of
267 standard to inhibit the cell viability by 20 % (IC₂₀) (Cañete and Diogène, 2008). Concentrations of CTX-
268 like compounds in microalgal extracts were estimated inferring the concentration from the standard
269 curve based on the viability of neuro-2a cells. The amounts of CTX-like compounds were expressed
270 as femtograms (fg) of CTX1B equiv. per cell. The limit of quantification (LOQ) was calculated as the
271 ratio of the LOD obtained with standard to the maximum concentration of microalgal extract used in
272 the assay with no matrix effect being observed. A matrix effect was considered when toxicity was
273 recorded in the neuro-2a cells after exposure to microalgal extracts without ouabain and veratridine
274 treatment (O/V-).

275 **2.6.3 MTX-like toxicity evaluation**

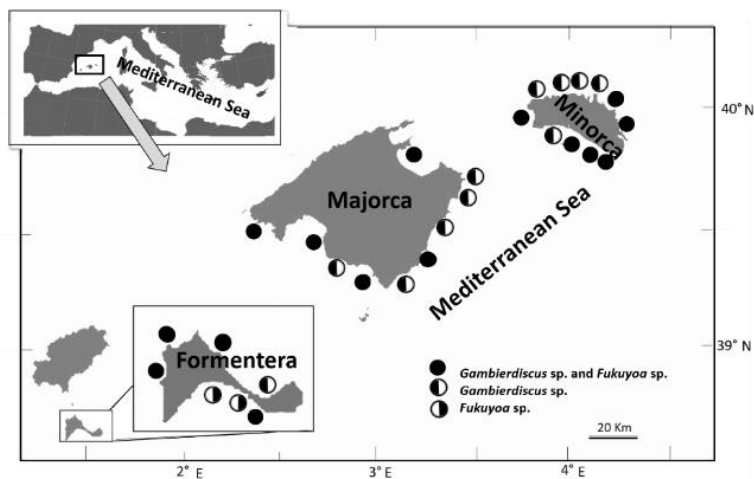
276 The MTX-like toxicity was evaluated qualitatively for 15 *G. australes* strains following the protocol
277 described by Caillaud et al. (2010). This assay is based on the inhibition of the toxic effect by the
278 addition of SKF96365, which is the 1-[2-(4-methoxyphenyl)-2-[3-(4-methoxyphenyl)propoxy]ethyl-
279 1H-imidazole hydrochloride to neuro-2a cells. SKF96365 blocks the voltage-gated Ca²⁺ channels
280 (VGCCs) (Singh et al., 2010) counteracting the increase of intracellular calcium levels caused by
281 compounds that target VGCCs.

282 **3. Results**

283 **3.1 Presence of *Gambierdiscus* and *Fukuyoa* genera in the western Mediterranean Sea.**

284 Presence of *Gambierdiscus* and *Fukuyoa* genera was assessed using samples from live and Lugol's
285 iodine preserved samples collected during 2016 to 2019. A total of 110 isolates from the genera
286 *Gambierdiscus* and *Fukuyoa* were obtained from the epiphytic samples and 26 isolates from the
287 epilithic samples. Epiphytic samples were obtained from macrophytes of the genera *Lobophora*,
288 *Cystoceira*, *Jania*, *Padina*, and *Dictyota*. Furthermore, other dinoflagellates co-occurred with the
289 *Gambierdiscus* and *Fukuyoa* genera, such as the genera *Prorocentrum*, *Coolia*, *Amphidinium* and
290 *Ostreopsis*. Figure 1 shows stations where the presence of the genera *Gambierdiscus* and *Fukuyoa*
291 were recorded in the Balearic Islands during the entire sampling period. Results of the presence of
292 the genera *Gambierdiscus* and *Fukuyoa* by sampling point and the environmental data (temperature,
293 pH, oxygen and salinity) are provided in supplementary Table 1. In Formentera, *Gambierdiscus* cells
294 were present in 5 out of 9 sampling stations, in low amounts in both samples (epilithic and epiphytic).
295 The presence was confirmed only in Lugol's iodine preserved samples; therefore, no live cells could
296 be isolated. In Majorca, *Gambierdiscus* cells were found both in epiphytic and epilithic samples,
297 although in Minorca cells were primarily found in the epiphytic samples. In Majorca and Minorca, in
298 2017 *Gambierdiscus* cells were present in all sampling stations with the exception of one site. Similar
299 results for both islands were obtained in 2018. In 2019 only Minorca was sampled, and in 2 out of 4
300 sampling stations *Gambierdiscus* cells were present.

301 The genus *Fukuyoa* was present in epiphytic and epilithic samples, in 6 out of 9 stations of
 302 Formentera. *Fukuyoa* cells were found at very low amounts and only in preserved samples, therefore
 303 as with the genus *Gambierdiscus*, no cells could be isolated in culture from Formentera. In 2017 in
 304 Majorca and Minorca, *Fukuyoa* cells were present in 5 out of 10 stations and 6 out of 9 stations
 305 respectively; while in 2018 cells were found in fewer stations: 4 out of 10 stations and 2 out of 9
 306 stations, respectively. In 2019 in Minorca, cells were not observed in any of 4 sampling stations.
 307 *Fukuyoa* isolates from Majorca and Minorca were obtained only from epiphytic samples. During the
 308 entire study period, *Fukuyoa* cells were concomitant with *Gambierdiscus* in 4 stations in Formentera,
 309 and in all stations in Majorca and Minorca.



310 Fig. 1. Presence of the *Gambierdiscus* and *Fukuyoa* genera in the sampling stations in the Balearic
 311 Islands (Mediterranean Sea) during 2016-2019.

312 3.2 Molecular characterization

313 Species level identification was performed for thirty-four *Gambierdiscus* and nine *Fukuyoa* isolates
 314 using the D8-D10 region (LSU) rDNA (Chinain et al. 1999, Litaker et al. 2009). Sequences were
 315 matched in GenBank using the BLAST sequence similarity searches (National Center for
 316 Biotechnology Information) and they scored the highest identity and similarity with *Gambierdiscus*
 317 *australes* and *Fukuyoa paulensis*. Moreover, further phylogenetic analyses confirmed the
 318 identifications (Fig.2).

319
320
321
322
323
324
325
326
327
328
329
330
331
332
333
334
335
336
337
338

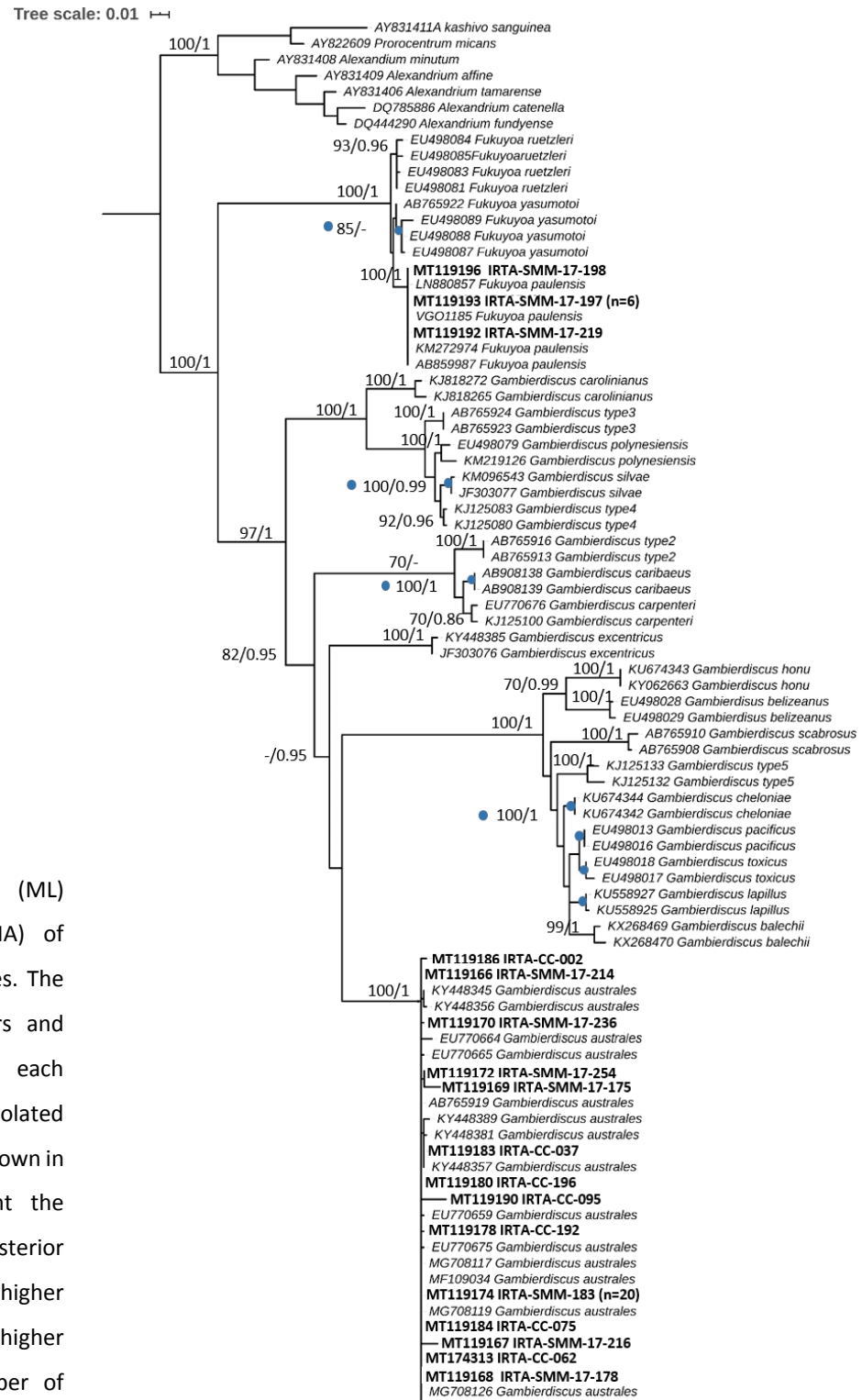
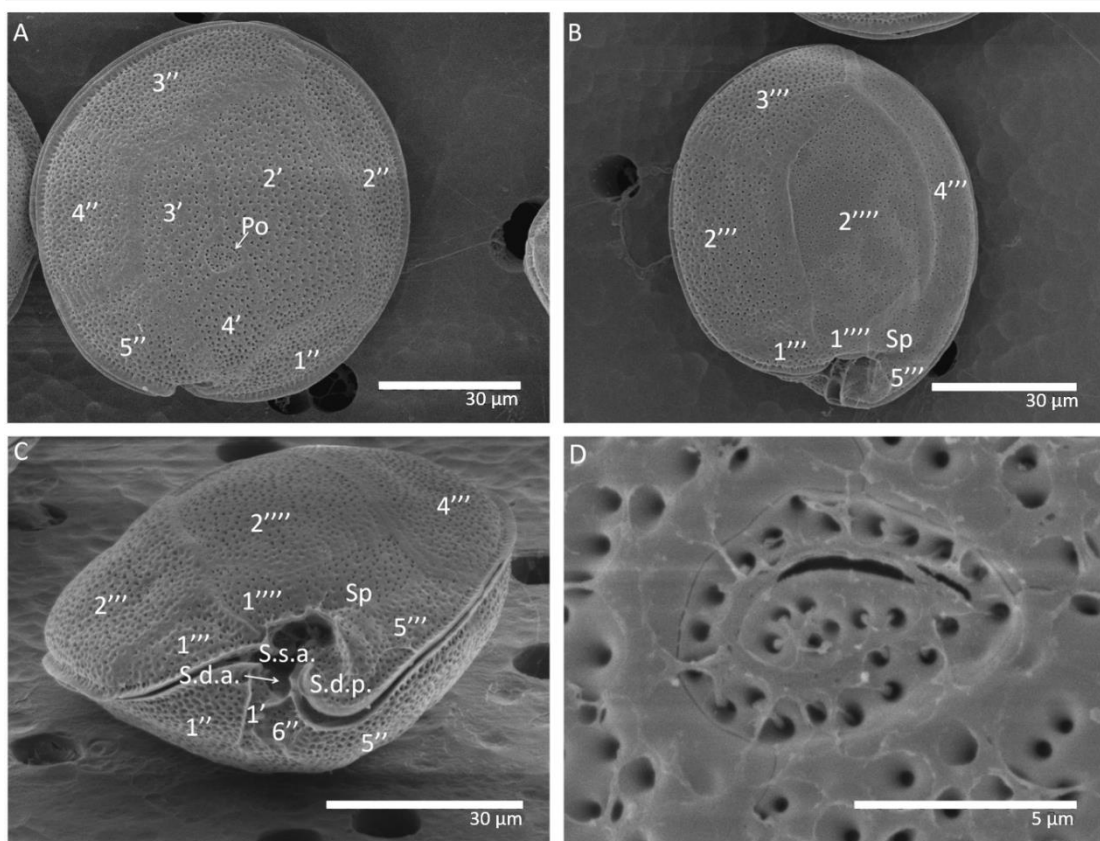


Fig. 2. Maximum likelihood (ML) phylogeny of D8-D10 LSU (rDNA) of *Gambierdiscus* and *Fukuyoa* species. The GenBank code accession numbers and species names are shown for each downloaded sequence. Strains isolated from samples from this study are shown in bold. Values at nodes represent the bootstrap values /Bayesian posterior probability. Only bootstraps values higher than 70 and posterior probabilities higher than 0.95 are shown. The number of clones (n) with the same haplotype is shown in parentheses.

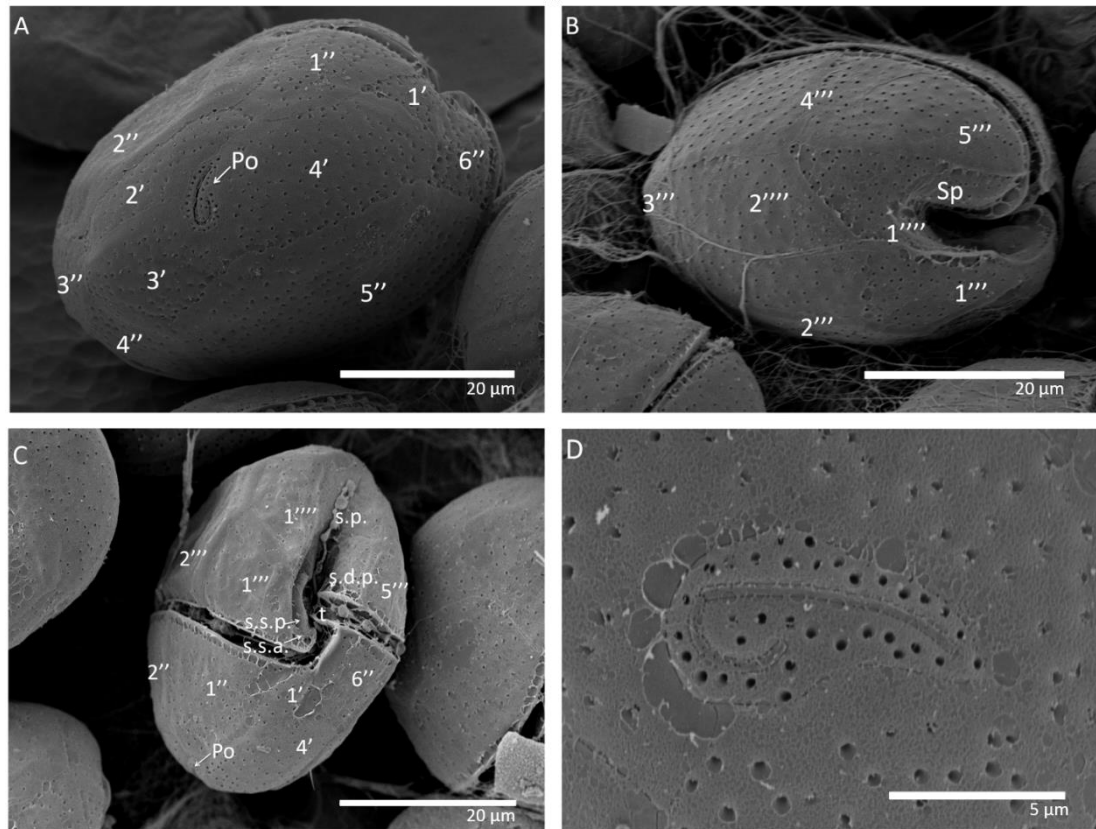
339 **3.3 Morphological characterization**

340 In the present study, cells of *G. australes* were anterior-posteriorly compressed showing a lenticular
 341 shape. Table 1 shows the measurements for *G. australes* and *F. paulensis* cells from the Balearic
 342 Islands in comparison to the measurements for these species retrieved from the literature. The thecal
 343 plate formula was: Po, 4', 0a, 6'', 6c, ?s, 5''', 0p, 2'''''. Fig. 3 shows a representative SEM photos for
 344 *G. australes*.

345 *F. paulensis* cells were globular presenting a lateral compression. Measurements are in Table 1. The
 346 thecal plate formula was Po, 4', 6'', 6c, ?s, 5''', 2'''''. Fig. 4 shows a representative SEM photos for *F.*
 347 *paulensis*.



348 **Fig. 3.** Images from SEM of *G. australes* (IRTA-SMM-17-253): apical (A), antapical (B), ventral (C)
 349 views, detail of Po plate and pores (D).



350 **Fig. 4.** SEM images of *F. paulensis* (IRTA-SMM-17-211): apical (A), antapical (B) and ventral (C)
 351 views and detail of Po plate and pores (D). (s. p.: sulcal posterior, s.s.p.: sulcal left posterior plate,
 352 s.d.p.: sulcal right posterior, s.s.a: sulcal left anterior plate).

353

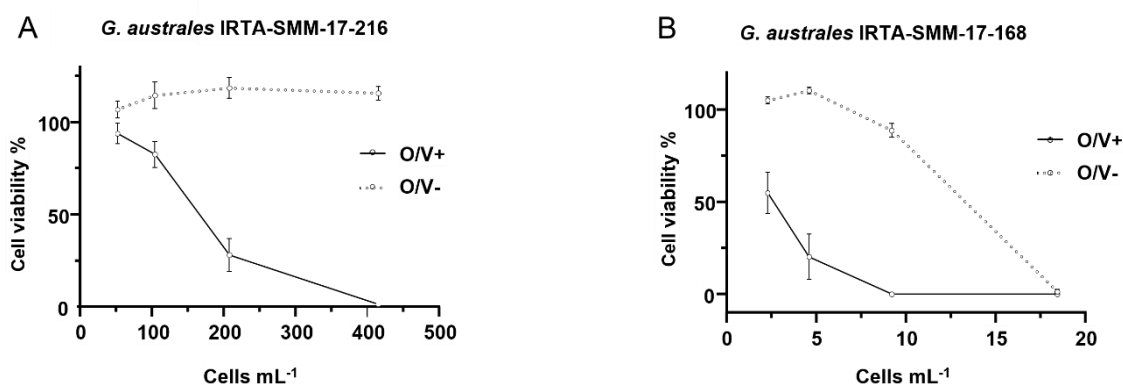
354 **3.4 Growth dynamics**

355 All the studied strains (three *G. australes* and one of *F. paulensis* strains) displayed a typical
 356 growth curve in batch culture conditions. No significant differences were observed among the
 357 replicates of the strains. Strains of *G. australes*: IRTA-SMM-17-162, IRTA-SMM-17-189 and IRTA-
 358 SMM-17-271 arrived at the stationary phase at the 25th, 22nd and 21st days. For *F. paulensis*,
 359 strain IRTA-SMM-17-209 reached the stationary phase at day 21st of culture. The growth curves
 360 are provided in supplementary material figure S1 and results of growth rates are shown in Table

361 2.

362 3.5 Evaluation of CTX-like and MTX-like toxicity

363 Exposure of neuro-2a cells to a CTX1B standard was nontoxic. As expected, addition of
 364 ouabain/veratridine (O/V+) showed a typical curve of CTX-like toxicity in neuro-2a cells with an
 365 average LOD of 0.45 ± 0.24 pg CTX1B equiv. \cdot mL⁻¹ and IC₅₀ of 1.21 ± 0.48 pg CTX1B equiv. \cdot mL⁻¹.
 366 The maximum concentration of microalgal extract that did not cause any toxicity in the absence
 367 of ouabain and veratridine ranged from 10 to 220 and from 40 to 450 cells equiv. \cdot mL⁻¹, for *G.*
 368 *australes* and *F. paulensis*, respectively.



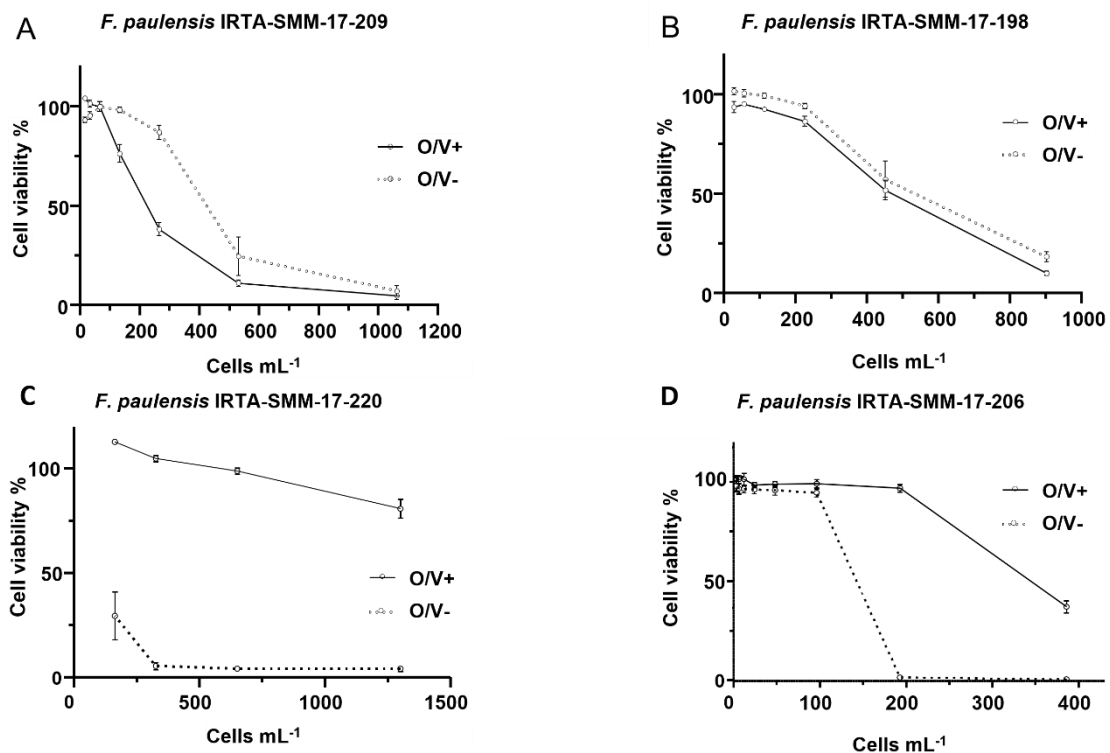
369

370 **Fig. 5.** Dose response curves obtained using neuro-2a CBA for *G. australes* extracts: IRTA-

371 SMM-17-216 (A), IRTA-SMM-17-168 (B). O/V+: neuro-2a cells exposed to microalgal extract
 372 with the ouabain and veratridine treatment. O/V-: neuro-2a cells exposed to microalgal extract
 373 without the ouabain and veratridine treatment. Each point is the mean of triplicates and bars
 374 represent the SD.

375 All *G. australes* extracts (n=21) presented CTX-like toxicity and toxicities ranged from 1.38 to 381
 376 fg CTX1B equiv. \cdot cell⁻¹ (Table 3). Figure 5 shows representative dose-response curves of the
 377 types of neuro-2a cell viability response for *G. australes*. Figure 5A corresponds to *G. australes*
 378 extract with low toxicity (IRTA-SMM-17-216). In the O/V+ conditions, the curve showed the
 379 typical dose-response curve of CTX-like toxicity with an estimated IC₅₀ of 150 cell equiv. \cdot mL⁻¹.

380 Figure 5B represents *G. australes* with high toxicity (IRTA-SMM-17-168) with an estimated IC_{50}
 381 of 2 cell equiv. $\cdot mL^{-1}$.



382

383 **Fig. 6.** Dose response curves obtained using neuro-2a CBA for *F. paulensis* extracts: IRTA- SMM-
 384 17-209 (A), IRTA-SMM-17-198 (B), IRTA-SMM-17-220 (C), IRTA-SMM-17-206 (D). O/V+: neuro-
 385 2a cells exposed to microalgal extract with the ouabain and veratridine treatment. O/V-: neuro-
 386 2a cells exposed to microalgal extract without the ouabain and veratridine treatment. Each point
 387 is the mean of triplicates and bars represent the SD.

388 Two strains of *F. paulensis* IRTA-SMM-17-209 and IRTA-SMM-17-211 showed CTX-like
 389 compounds and the remaining four extracts did not show CTX-like toxicity (Table 3). Figure 6A
 390 corresponds to *F. paulensis* extract (IRTA-SMM-17-209). Cell exposure to this extract at <260
 391 cells equiv. $\cdot mL^{-1}$ under O/V- conditions, resulted in no significant toxicity, while at >120 cells
 392 equiv. $\cdot mL^{-1}$ in the O/V+ conditions significant toxicity was recorded, indicating a CTX-like effect.
 393 Figure 6B corresponds to *F. paulensis* extract (IRTA-SMM-17-198). Under both conditions with
 394 and without (O/V), the cell inhibition was significant, therefore no conclusion could be drawn in

395 reference to CTX-like toxicity. Two *F. paulensis* extracts (IRTA-SMM-17-206 and IRTA-SMM-17-
 396 220) caused cell mortality of neuro-2a in the absence of O/V (Fig. 6C, 6D). Nonetheless, under
 397 O/V+ conditions, the toxicity of these extracts was decreased, and this is a novel toxicity pattern
 398 described for this genus.

399 In order to confirm the presence of MTX-like toxicity, neuro-2a cells were exposed to microalgal
 400 extracts in the presence of SKF96365. Twelve out of fifteen *Gambierdiscus* strains showed
 401 recovery of the cell viability when SKF96365 was added (Table 3). Figure 7 shows representative
 402 dose response curve of a *G. australes* extract with and without SKF96365. On the contrary, two
 403 *Fukuyoa* strains did not show recovery of the cell viability when SKF96365 was added.

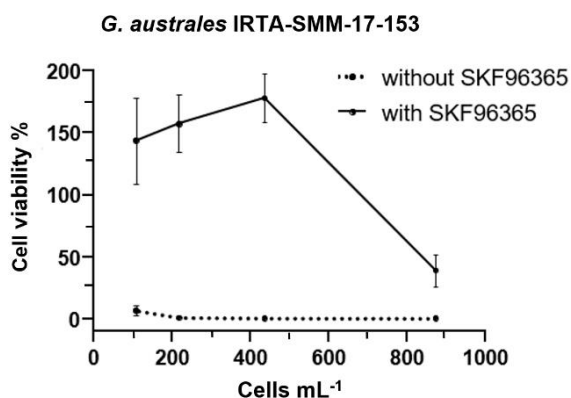


Fig 7. Dose response curve obtained using neuro-2a CBA with the *G. australes* extract (IRTA-SMM-17-153). Without SKF96365: neuro-2a cells exposed to extract in the absence of SKF96365; with SKF96365: neuro-2a cells exposed to toxin extracts in the presence of SKF96365. Each point is the mean of triplicates and the bars represent the SD.

404

405 4. Discussion

406 The presence of the genus *Gambierdiscus* in the Eastern Mediterranean Sea was reported in
 407 2003 (Aligizaki & Nikolaidis, 2008). Reported species include *G. carolinianus* (Holland et al.,
 408 2013), *Gambierdiscus* sp., *G. cf. belizeanus* and *G. silvae* (Aligizaki et al., 2018). *G. australes* was
 409 detected later in the Balearic Islands, as presented in a brief communication (Tudó et al., 2018).
 410 The first detection of the genus *Fukuyoa* was in 2016 in the Western Mediterranean Sea (Laza-
 411 Martínez et al., 2016) and in 2018 in the Eastern Mediterranean Sea (Aligizaki et al., 2018).

412 To the best of our knowledge, the Balearic Islands is the location with the highest latitude
 413 worldwide, where the *Gambierdiscus* genus has been detected, specifically at 40.06 ° N. In the

414 present study, the presence of the genus *Gambierdiscus* over large areas along the coasts of the
415 Balearic Islands and the recurrence at some stations over three years suggests that this genus is
416 well-established in the archipelago. However, the genus *Fukuyoa* was identified in 2017 and
417 2018, but not in 2019. Though the absence in 2019 should take into account that only four
418 stations were sampled that year.

419 Water temperature influences on the growth and cell abundance of microalgae and can predict
420 latitudinal distribution. *Gambierdiscus* species show different thermal limits, and distinct
421 optimal temperatures (Kibler et al., 2012). Other variables such as salinity and irradiance can
422 play an important role in the species distribution, though, their limits may be common for
423 several species (Kibler et al., 2012). During the sampling dates, for Formentera, Minorca and
424 Majorca water temperatures (22.8 and 27.2 °C) were close to optimal temperatures for
425 *Gambierdiscus* species. Generally, for *Gambierdiscus* spp. the optimal temperature range is
426 between 23 and 29 °C and the survival below 15 °C in laboratory conditions is rare (Kibler et al.,
427 2012; Xu et al., 2016). *G. australes* is one of the most cryo-tolerant species in the genus; its
428 optimal temperature for growth is relatively low, at 25 °C (Tester et al., 2020). Besides, *G.*
429 *australes* cells from the Canary Islands showed the ability to stay alive with no growth for six
430 months at 15 °C, and they resumed growth when the temperature arose to 17 °C (personal
431 communication by Dr. Isabel Bravo <https://ciguateravgo.es/>). This thermo-physiologic
432 characteristic of *G. australes* could confer upon this species the ability to persist in the Balearic
433 Islands in wintertime when the water temperature drops at 13 °C. Regarding *Fukuyoa* spp., any
434 literature of the optimal temperature for growth is scarce. Nonetheless, *F. paulensis* (classified
435 previously as *G. yasumotoi*) was recorded in New Zealand (Rhodes et al., 2014a), where the
436 water temperature oscillates between 14 and 23 °C. Additionally, one strain (Dn135EHU) from
437 the Balearic Islands showed the formation of cysts (Laza-Martínez et al., 2016), and this could
438 favour the species survival for long periods at low temperatures.

439 In previous studies of microalgae samples collected in Majorca in 1997-1998 (Vila et al., 2001a),
440 2001 (Penna et al., 2005) and 2011 (Laza-Martínez et al., 2016) no cells of the genera
441 *Gambierdiscus* or *Fukuyoa* had been detected. In addition, during 2005-2006 an exhaustive
442 sampling was conducted to characterize the phytoplankton communities from 1 to 15 m depth
443 in the entire Balearic Archipelago (Puigserver et al., 2008), and cells of the genera *Gambierdiscus*
444 or *Fukuyoa* were not detected. Although *Gambierdiscus* and *Fukuyoa* are mainly benthic, and
445 Puigserver et al. (2008) was focused on phytoplankton, free-swimming cells could have been
446 observed as was described in Parsons et al. (2011). The recent findings of cells of these genera
447 in the Balearic Islands could be explained by an intense and specific sampling design for benthic
448 species. Although cell abundance was not evaluated in the present work, the recent detection
449 of these genera might be a result of an increase in abundance of endemic populations. The
450 populations could be influenced by climate change (Aligizaki et al., 2008; Kibler et al., 2015;
451 Llewellyn, 2010, Larsson et al., 2019), which in the Mediterranean Sea, is expected to cause an
452 increase in abundance of thermo-tolerant species and a decrease or disappearance of cold-
453 tolerant stenothermal species (Lejeusne et al., 2009). In addition to regional temperature
454 increase as potential cause to changes in microalgal populations, other factors such as storms
455 or anthropogenic activities in coastal regions could be involved. It has been suggested that
456 expansion of benthic dinoflagellates could be attributed to an increase of turf algal mats
457 covering substrates due to environmental changes (storms, currents, acidification) (Kohler and
458 Kohler 1992; Rongo and van Woesik, 2013; Turquet et al., 2001), but also to the degradation of
459 the marine environment directly associated to human activities such as bottom dredging for the
460 creation of port structures and other forms of coastal embayments, drag-netting, pollution and
461 over-exploitation of natural resources (Parsons and Preskitt, 2007; Skinner et al., 2013; Vila et
462 al., 2001c; Villareal et al., 2007). The environment of the Balearic Islands, most specifically the
463 coastal areas, has suffered extreme pressures from tourism since the 1960s (Garín-Muñoz and
464 Montero-Martín, 2007). The impact of tourism has caused a clear degradation of the coast by

465 increasing port structures, disturbing the coastal sediments and increasing the eutrophication
466 (Puigserver et al., 2002). In addition, meadows of *Posidonia oceanica* in the Balearic Islands are
467 in decline in favour of colonisation of turf algal mats (Ballesteros et al., 2007; Duarte et al., 2009).
468 It has been suggested that reduction of these disturbances should not be expected in the coming
469 years (Duarte et al., 2009; Garín-Muñoz and Montero-Martín, 2007), so these factors could
470 favour further increase in *Gambierdiscus* populations.

471 Another explanation for the presence of *Gambierdiscus* cells, may be new colonisations from
472 other regions. For some benthic species of toxic dinoflagellates, the new colonisations may be
473 associated with translocations of organisms by ballast waters (Hallegraeff, 2015). In fact, it has
474 been suggested that *Alexandrium pacificum* (previously identified as *Alexandrium catenella*), a
475 species described as non-native in the Mediterranean Sea, has been introduced by ballast waters
476 of cargo vessels (Vila et al. 2001b, 2001c). It is well reported that the Eastern Mediterranean
477 Basin is suffering a large-scale invasion of tropical and subtropical species. At the moment, more
478 than 700 species of organisms have been identified as having come from the Red Sea through
479 the Suez channel (Zenetos et al., 2012). However, from the genera *Gambierdiscus* and *Fukuyoa*,
480 the only species reported in the Red Sea is *G. belizeanus* (Catania et al., 2017) and *F. yasumotoi*
481 (Saburova et al., 2013). Therefore, the phenomenon of species translocation from the Red Sea
482 to the Mediterranean Sea may not explain the current situation for these species. Considering
483 the possibility that *Gambierdiscus* cells reached the Balearic Islands from the Atlantic Ocean, the
484 genetic information provided by the D8-D10 region (LSU rDNA) shows that *G. australes*
485 populations from the Balearic Islands and the Canary Archipelago are identical. Although this
486 information may establish a link between these populations, more molecular markers should be
487 analysed to determine the relationship between populations in these two areas. Population
488 genetics and phylogeographic studies of these species have to be considered in future studies
489 because they can help to identify the source of populations and reveal expansion patterns, and
490 mechanisms of transfer (Sakai et al., 2001).

491 *Gambierdiscus australes* cells from the Balearic Islands show morphological similarities to other
492 *G. australes* described in previous studies (Table 1). Cell size (D and W) from the present study
493 are partially consistent with the range of the first description of *G. australes* for the strain RAV-
494 92 in the Pacific Ocean (Chinain et al., 1999). Later, strains RAV-92 and NOAA2 were measured
495 by Litaker et al., (2009) and their minimum extreme values of D and W were almost the same as
496 in the current study. Cell morphology can change by natural factors, but also, over time of
497 cultures in laboratory conditions (Rhodes et al., 2014b). The maximum values of D and W for
498 *G. australes* were described for strains isolated from the Canary Islands (Atlantic Ocean) in Bravo
499 et al. (2019). Values of Bravo et al. (2019) and the present work show larger sizes than those of
500 Rhodes et al. (2014b), Chinain et al. (1999) and Litaker et al. (2009).

501 Regarding *Fukuyoa* isolates, the average L and D of cell size in the current study are inside the
502 ranges of previous studies performed in the Mediterranean Sea, the Atlantic and the Pacific
503 Ocean (Gómez et al., 2015, Laza-Martínez et al., 2016, Rhodes et al., 2014a) (Table 1). However,
504 the lowest values for L and W in the present work are smaller than in the previous studies.

505 The maximum cell yield for *G. australes* cultures in the present study of growth dynamics and
506 toxicity was 2288 to 2274 cells · mL⁻¹, respectively. These values are lower than for *G. australes*
507 strains from the North Atlantic Ocean (Reverté et al., 2018), where the maximum cell yield was
508 4470 cells · mL⁻¹. Such differences may be attributed to the differences in the culturing
509 conditions. In both works, cells were cultivated at the same temperature and medium, but in
510 Reverté et al. (2018) photon irradiance was lower, a pump supplied the aeration and the vessel
511 was a 3L round-bottom flask.

512 In the current work, growth rates for *G. australes* were lower than the rates of *G. australes*
513 strains from the Atlantic Ocean reported by Reverté et al. (2018), which ranged from 0.20 to
514 0.39 div. · day⁻¹, and they are similar to the rates for *G. australes* strains, from the Pacific Ocean
515 described before: 0.12 - 0.19 div. · day⁻¹ in Chinain et al. (2010) and 0.149 ± 0.006 div. · day⁻¹ in

516 Pisapia et al. (2017). The growth rate in the *Gambierdiscus* genus has been reported to be in the
517 range of 0.01 to 0.55 div. · day⁻¹ (Xu et al., 2016; Whitters, 1981). Some studies for the genus
518 *Gambierdiscus* link high division rates to high toxin production per cell (Chinain et al., 2010,
519 Litaker et al., 2017; Pisapia et al., 2017, Reverté et al., 2018), but in the present study, this
520 relation was not studied.

521 Regarding *F. paulensis* growth, there was high variability of maximum cell yield among the
522 strains from the current study. The maximum cell yield in the growth dynamics study was 1004
523 cells · mL⁻¹ (Table 2), and for the CTX-like toxicity study, values ranged between 333 and 6636
524 cells · mL⁻¹ (Table 3). These yields were much lower than those achieved at the stationary phase
525 by Laza-Martínez et al. (2016) of 14.800 cells · mL⁻¹. In Laza-Martínez et al. (2016), strains were
526 cultured in culture plastic flasks, with f/4 medium with selenium (Guillard and Ryther, 1962) and
527 salinity was adjusted at 35. Besides, cells were maintained at 25 °C and irradiance of 50-100
528 μmol · m⁻² · s⁻¹. Regarding growth rates, the present study provides the first data for *F. paulensis*
529 with 0.24 ± 0.06 div. · cell⁻¹. Within the *Fukuyoa* genus, *F. ruetzleri* (previously *G. ruetzleri*)
530 showed growth rates of 0.17, 0.18 and 0.35 div · cell⁻¹ in Litaker et al. (2017), Pisapia et al. (2017),
531 and Kibler et al. (2012), respectively.

532 In the current paper, *G. australes* strains presented CTX-like activity with quantifications ranging
533 between 1.4 and 380 fg CTX1B equiv. · cell⁻¹. These quantifications are low compared to values
534 for *G. australes* from the Atlantic Ocean reported by Reverté et al. (2018), where values ranged
535 from 200 to 697 fg CTX1B equiv. · cell⁻¹. In both works, strains were acclimated for one year, but
536 as it has been mentioned before, they were cultured in different culturing conditions. Therefore,
537 dissimilar toxin production could be caused by distinct culturing conditions. In contrast, the CTX-
538 like activity was similar to other *G. australes* strains from the Atlantic Ocean (31-107 fg CTX1B
539 equiv. · cell⁻¹) and from the Pacific Ocean (40 fg CTX1B equiv. · cell⁻¹) reported by Rossignoli et
540 al. (2020) and Rhodes et al. (2017), respectively.

541 Among *Gambierdiscus* species, *G. australes* has intermediate CTX-like toxicity. For instance, by
542 standard mouse bioassay (MBA), *G. australes* extracts presented lower toxicity than *G. pacificus*
543 and *G. polynesiensis* (Chinain et al., 1999), with the latter being the most toxic species in the
544 genus. Furthermore, in Chinain et al. (2010), the CTX-like response for strains from the Pacific
545 Ocean was similar to *G. toxicus*, and 100-fold lower than in *G. polynesiensis*. Moreover, in Pisapia
546 et al. (2017), the CTX-like toxicity of ten strains was evaluated by neuro-2a CBA, and three *G.*
547 *australes* strains (two from the Atlantic and one from the Pacific Ocean) were placed in the
548 seventh-place, near the bottom of the scale.

549 Despite several unsuccessful attempts to confirm toxicity in *Gambierdiscus* spp. (Larsson et al.,
550 2018), CTXs have not been confirmed for most *Gambierdiscus* spp., except for *G. australes*
551 (Roeder et al., 2010), *G. pacificus* (Caillaud et al., 2011), *G. polynesiensis* (Chinain et al., 2010)
552 and *G. excentricus* (Paz et al., 2011). A putative CTX (2,3-dihydroxy P-CTX-3C) was detected by
553 liquid chromatography-mass spectrometry (LC-MS/MS) in only one *G. australes* strain (CCMP
554 1653) from Hawaii (Pacific Ocean), at exponential phase (Roeder et al., 2010). In fact,
555 *Gambierdiscus* species are common producers of MTXs and *G. australes* is one of the top
556 producers (Munday et al., 2017). MTX1 and MTX3 were detected by LC-MS/MS in *G. australes*
557 by Munday et al. (2017). Moreover, in Rhodes et al. (2017), LC-MS/MS confirmed the presence
558 of MTX1 in all tested *G. australes* strains. This is in accordance with the results for the present
559 study in which almost all strains of *G. australes* (12 out of 15) presented MTX-like activity.
560 Nonetheless, a recent study including *G. australes* cultures from the present work, (IRTA-SMM-
561 17-162, IRTA-SMM-17-164, IRTA-SMM-17-189, IRTA-SMM-17-244, IRTA-SMM-17-253, IRTA-
562 SMM-17-271) were analysed by liquid chromatography coupled to low and high resolution mass
563 spectrometry (LC-MS/MS) and (LC-HRMS), and MTX1, desulfo-MTX1 and didehydro-34 desulfo-
564 MTX1 were not detected. By contrast, 44-methylgambierone (MTX3) was present in all of these
565 strains (Estevez et al., 2020).

566 The MTX family, previously included only molecules with a molecular weight of more than 3000
567 Da and having no activity on VGSCs (Yokoyama et al., 1988). Recently, 44-methylgambierone has
568 been found in *G. belizeanus* and *G. australes* (Boente-Juncal et al., 2019; Murray et al., 2019).
569 This molecule was previously defined as MTX3 (Holmes and Lewis, 1994). Nonetheless, it is a
570 molecule of 1060 Da which presents structural differences as compared to the previous MTXs
571 (Holmes and Lewis, 1994) and shows CTX-like activity more than MTX-activity (Boente-Juncal et
572 al., 2019). In human cortical neurons, 44-methylgambierone showed no signals of cell mortality
573 at 0.01 to 20 nM for five days, whereas at 0.1 nM of MTX1 significant cell death was observed
574 in 2h (Boente-Juncal et al., 2019). Furthermore, in human neuroblastoma cells, after 24h of
575 exposure of cells to MTX3 at 10 to 50 nM cells did not show signs of toxicity, while with MTX1
576 at 0.1 nM, complete cell death was observed (Boente-Juncal et al., 2019). Hence, including 44-
577 methylgambierone in the MTX group may lead to confusion on the role of the rest of MTXs in
578 CP. Given that only one strain of *G. australes* strain produced a CTX analogue (Rhoeder et al.,
579 2010), and the effects of 44-methylgambierone in neuro-2a cells could be similar to effects of
580 CTXs, the CTX-like toxicity of the *G. australes* extracts of the present study could be potentially
581 attributed to the effect of 44-methylgambierone. Even so, 44-methylgambierone exhibited very
582 low toxicity by MBA, hence it is unlikely it contributes to CP (Murray et al., 2020).

583 Concerning *F. paulensis*, the CTX-like toxicities of the current study ranged from 8 to 16 fg CTX1B
584 equiv. · cell⁻¹. These toxicities were low in comparison to the *G. australes* strains. Previously, one
585 *F. paulensis* strain (Dn35EHU) from the Balearic Islands presented low CTX-like toxicity by MBA
586 (Laza-Martínez et al., 2016). In the same study, for the same strain, traces of 54-deoxyCTX1B
587 and gambieric acid A (GA A) were detected by LC-HRMS. Recently, Estevez et. al (2020) detected
588 44-methylgambierone for the *F. paulensis* strain (IRTA-SMM-17-209), which is the same strain
589 used in the current study. This is in accordance with the results of Rhodes et al. (2014a) and
590 Larsson et al. (2019), which detected 44-methylgambierone in *F. paulensis* from the Pacific

591 Ocean. Therefore, like *G. australes*, the CTX-like toxicity of *F. paulensis* in neuro-2a CBA could be
592 explained by the presence of CTX analogues and the 44-methylgambierone.

593 *F. paulensis* presents low toxicity in comparison to other *Fukuyoa* species. Litaker et al. (2017)
594 detected CTX-like toxicity in three *F. ruetzleri* strains by neuro-2a CBA with an average of 24.50
595 and 6.50 fg CTX3 equiv. · cell⁻¹. Like *G. australes*, it is suggested that some strains of *F. paulensis*
596 are non-CTX-producers because no signal of CTX-like toxicity and no CTX-analogues were found
597 at early stationary phase for a cultured strain (VGO1185) from Brazil (Atlantic Ocean) by neuro-
598 2a CBA (Gómez et al., 2015). Moreover, *F. paulensis* (previously *G. yasumotoi*) CAWD210 from
599 New Zealand did not exhibit CTX-like toxicity by sea urchin embryo assay (SUEA) (Rhodes et al.,
600 2014a). That is in concordance with the results of the present study and the increase in viability
601 observed when neuro-2a cells were exposed to extracts from *F. paulensis*. Laza-Martínez et al.
602 (2016) detected MTX-like activity by MBA. However, in the present work, the MTX-like activity
603 for *F. paulensis* was not detectable (n=2). Furthermore, there is no confirmation of MTX1,
604 desulfo-MTX1 and didehydro-34 desulfo-MTX1 in *F. paulensis* strain (IRTA-SMM-17-209) by
605 analytical methods (Estevez et al. 2020).

606 Toxicities of several strains of *Gambierdiscus* from the Pacific have been largely studied, but
607 information about the strains from the Mediterranean Sea is scarce despite the increasing
608 identification of species in recent decades. To the best of our knowledge, the presence of CTX-
609 like toxicity in *Gambierdiscus* strains has only been evaluated from the eastern Mediterranean
610 region using three strains of *G. carolinianus*, *G. silvae* and *Gambierdiscus* sp.; all of them
611 analysed by neuro-2a CBA. The *G. carolinianus* strain showed CTX-like activity in low quantities
612 (< 4 fg CTX3C equiv. · cell⁻¹) (Pisapia et al., 2017), the *G. silvae* showed high CTX-like toxicity and
613 the putative new species *Gambierdiscus* sp. exhibited low CTX-like activity (Aligizaki et al., 2018).
614 The demonstration of CTX-like toxicity in strains in the Balearic Islands, and the fact that no

615 evidence of ciguateric fish or CP has occurred in this area, could suggest that these populations
616 are relatively new residents or that the densities of the populations are probably low.

617 **5. Conclusions**

618 *Fukuyoa* and *Gambierdiscus* cells found in samples from the Balearic Islands from 2016 to 2019
619 have been identified as *F. paulensis* and *G. australes*. These two species seem to be well-
620 established in the area. Considering the other studies, CTX-toxicity exhibited by most of the *G.*
621 *australes* and *F. paulensis* strains was low. However, one strain of *G. australes* (IRTA-SMM-17-
622 168) was classified as a very high producer in comparison to previous studies. In addition, it is
623 not possible to discard that some cells from the Balearic Islands could be a high CTX-producers
624 and could be associated with distinct seasonality. Even though CP cases have not yet been
625 confirmed in the Mediterranean, the CTX-like toxicity present in the strains of *G. australes* and
626 *F. paulensis* from the Balearic Islands may indicate that potential future cases of CP should not
627 be dismissed. There is a clear need for continued studies and monitoring of benthic
628 dinoflagellates in the region.

629 **Conflict of interest**

630 The authors declare that there is no conflict of interest.

631 **Authors contribution**

632 Conceptualization A.T. (Àngels Tudó), M.C. (Mònica Campàs), M.F. (Margarita fernández-
633 Tejedor) and J.D. (Jorge Diogène); methodology A.T., A.T.F. (Anna Toldrà), M.R. (María Rey) and
634 I.T. (Irene Todolí); data curation A.T., K.A. (Karl B. Andree), M.F., J.D. formal analysis A.T., K.A.,
635 M.F., M.C., F.S. (Francesc X. Sureda) and J.D.; writing original draft preparation A.T., M.C., A.T.F
636 and J.D. ; writing, review and editing A.T., A.T.F., M.R., M.C., K.A., M.F., F.S. and J.D. All authors
637 have read and agreed to the published version of the manuscript.

638 **Acknowledgements**

639 The authors acknowledge José Luis Costa and Vanessa Castan for their assistance in the
640 samplings and to Ferran Pellisé for sampling in Menorca. We would like to thank to José Manuel
641 Fortuño from SEM facilities from Institut Ciències del Mar (CSIC). The authors acknowledge the
642 financial support from the European Food Safety Authority (EFSA) through the EUROCIGUA
643 project (GP/EFSA/AFSCO/2015/03) and the Ministerio de Ciencia, Innovación y Universidades
644 (MICINN), the Agencia Estatal de Investigación (AEI) and the Fondo Europeo de Desarrollo
645 Regional (FEDER) through the CIGUASENSING project (BIO2017-87946-C2-2-R). The authors also
646 acknowledge support from CERCA Programme / Generalitat de Catalunya and A. Toldrà and À.
647 Tudó acknowledge IRTA-URV-Santander for their respective PhD grants (2015 PMF-PIPF-67 and
648 2016 PMF-PIPF-74).

649 **References**

- 650 Adachi, R., Fukuyo, Y., 1979. The thecal structure of a marine toxic dinoflagellate *Gambierdiscus*
651 *toxicus* gen. et spec. nov. collected in a ciguatera-endemic area. Bull. Japanese Soc. Sci.
652 Fish. 45, 67–71.
- 653 Aligizaki, K., Iliadou, M., Kappas, I., Arsenakis, M., 2018. Is the eastern Mediterranean a
654 “*Gambierdiscus* biodiversity hotspot”? New data from Greece and Cyprus., 18th
655 International Conference on Harmful Algae. Nantes, France.
- 656 Aligizaki, K., Nikolaidis, G., Fraga, S., 2008. Is *Gambierdiscus* spreading to new areas? Harmful
657 Algae News 36, 6–7.
- 658 Bagnis, R., 1993. Algal Toxins in Seafood and Drinking Water. Acad. Press 105–115.
- 659 Bagnis, R., Chanteau, S., Chungue, E., Hurtel, J.M., Yasumoto, T., Inoue, A., 1980. Origins of
660 ciguatera fish poisoning: a new dinoflagellate, *Gambierdiscus toxicus* Adachi and Fukuyo,
661 definitively involved as a causal agent. Toxicon. 18, 199-208.
- 662 Balech, E., 1989. Redescription of *Alexandrium minutum* Halim (Dinophyceae) type species of

- 663 the genus *Alexandrium*. *Phycologia*. 206-211.
- 664 Ballesteros, E., Cebrian, E., Alcoverro, T., 2007. Mortality of shoots of *Posidonia oceanica*
665 following meadow invasion by the red alga *Lophocladia lallemandii*. *Bot. Mar.* 50, 8–13.
- 666 Bentur, Y., Spanier, E., 2007. Ciguatoxin-like substances in edible fish on the eastern
667 Mediterranean. *Clin. Toxicol.* 45, 695–700.
- 668 Bienfang, P., DeFelice, S., Dowing, A., 2011. Quantitative evaluation of commercially available
669 test Kit for Ciguatera. *Food Nutr. Sci.* 2, 594–598.
- 670 Boada, L.D., Zumbado, M., Luzardo, O.P., Almeida-González, M., Plakas, S.M., Granade, H.R.,
671 Abraham, A., Jester, E.L.E., Dickey, R.W., 2010. Ciguatera fish poisoning on the West Africa
672 Coast: An emerging risk in the Canary Islands (Spain). *Toxicon.* 56, 1516-1519.
- 673 Boente-Juncal, A., Álvarez, M., Antelo, Á., Rodríguez, I., Calabro, K., Vale, C., Thomas, O.P.,
674 Botana, L.M., 2019. Structure elucidation and biological evaluation of maitotoxin-3, a
675 homologue of gambierone, from *Gambierdiscus belizeanus*. *Toxins (Basel)*. 11.
- 676 Bomber, J.W., Tindall, D.R., Miller, D.M., 1989. Genetic variability in toxin potencies among
677 seventeen clones of *Gambierdiscus toxicus* (Dinophyceae). *J. Phycol.* 25.
- 678 Bravo, I., Rodriguez, F., Ramilo, I., Rial, P., Fraga, S., 2019. Ciguatera-causing dinoflagellate
679 *Gambierdiscus* spp. (Dinophyceae) in a subtropical region of North Atlantic Ocean (Canary
680 Islands): Morphological characterization and biogeography. *Toxins (Basel)*. 11.
- 681 Bravo, J., Cabrera Suárez, F., Ramírez, A.S., Acosta, F., 2015. Ciguatera, an Emerging Human
682 Poisoning in Europe. *J. Aquac. Mar. Biol.* 3, 1–6.
- 683 Caillaud, A., de la Iglesia, P., Barber, E., Eixarch, H., Mohammad-Noor, N., Yasumoto, T., Diogène,
684 J., 2011. Monitoring of dissolved ciguatoxin and maitotoxin using solid-phase adsorption
685 toxin tracking devices: Application *Gambierdiscus pacificus* in culture. *Harmful Algae* 10,

- 686 433–446.
- 687 Caillaud, A., Yasumoto, T., Diogène, J., 2010. Detection and quantification of maitotoxin-like
688 compounds using a neuroblastoma (Neuro-2a) cell-based assay. Application to the
689 screening of maitotoxin-like compounds in *Gambierdiscus* spp. *Toxicon* 56, 36–44.
- 690 Cañete, E., Diogène, J., 2008. Comparative study of the use of neuroblastoma cells (Neuro-2a)
691 and neuroblastoma × glioma hybrid cells (NG108-15) for the toxic effect quantification of
692 marine toxins. *Toxicon* 52, 541–550.
- 693 Catania, D., Richlen, M.L., Mak, Y.L., Morton, S.L., Laban, E.H., Xu, Y., Anderson, D.M., Chan, L.L.,
694 Berumen, M.L., 2017. The prevalence of benthic dinoflagellates associated with ciguatera
695 fish poisoning in the central Red Sea. *Harmful Algae* 68, 206–216.
- 696 Catterall, W.A., 1986. Molecular Properties of Voltage-Sensitive Sodium Channels. New Insights
697 into Cell Membr. Transp. Process. 3–20.
- 698 Catterall, W.A., Nirenberg, M., 1973. Sodium uptake associated with activation of action
699 potential ionophores of cultured neuroblastoma and muscle cells. *Proc. Natl. Acad. Sci. U.*
700 *S. A.* 70, 3759–3763.
- 701 Chan, T.Y.K., 2016. Characteristic features and contributory factors in fatal ciguatera fish
702 poisoning-implications for prevention and public education. *Am. J. Trop. Med. Hyg.*
- 703 Chinain, M., Darius, H.T., Ung, A., Cruchet, P., Wang, Z., Ponton, D., Laurent, D., Pauillac, S., 2010.
704 Growth and toxin production in the ciguatera-causing dinoflagellate *Gambierdiscus*
705 *polynesiensis* (Dinophyceae) in culture. *Toxicon* 56, 739–750.
- 706 Chinain, M., Gatti, C.M., Roué, M., Darius, H.T., 2019. Ciguatera poisoning in French Polynesia:
707 insights into the novel trends of an ancient disease. *New Microbes New Infect.*
- 708 Darriba, D., Taboada, G.L., Doallo, R., Posada, D., 2012. jModelTest 2: more models, new

- 709 heuristics and high-performance computing Europe PMC Funders Group. Nat. Methods 9,
710 772.
- 711 De Haro, L., Pommier, P., 2006. Hallucinatory fish poisoning (ichthyallyeinotoxism): Two case
712 reports from the Western Mediterranean and literature review. Clin. Toxicol. 44, 185–188.
- 713 Diogène, J., Reverté, L., Rambla-Alegre, M., Del Río, V., De La Iglesia, P., Campàs, M., Palacios,
714 O., Flores, C., Caixach, J., Ralijaona, C., Razanajatovo, I., Pirog, A., Magalon, H., Arnich, N.,
715 Turquet, J., 2017. Identification of ciguatoxins in a shark involved in a fatal food poisoning
716 in the Indian Ocean. Sci. Rep. 7:8240.
- 717 Duarte, C.M., Culbertson, J., Dennison, W.C., Fulweiler, R.W., Hughes, T., Kinney, E.L., Marbà, N.,
718 Nixon, S., Peacock, E.E., Smith, S., Valiela, I., 2009. Global Loss of Coastal Habitats Rates,
719 Causes and Consequences, 1st ed. BBVA Foundation, Bilbao.
- 720 Estevez, P., Sibat, M., Leao, J.M., Tudó, À., Rambla-Alegre, M., Aligizaki, K., Gago-Martinez, A.,
721 Diogène, J., Hess, P., 2020. Use of Mass Spectrometry to determine the Diversity of Toxins
722 Produced by *Gambierdiscus* and *Fukuyoa* Species from Balearic Islands and Crete
723 (Mediterranean Sea) and the Canary Islands (Northeast Atlantic). Toxins (Basel). 1–22.
- 724 Fernández-Zabala, J., Tuya, F., Amorim, A., Soler-Onís, E., 2019. Benthic dinoflagellates: Testing
725 the reliability of the artificial substrate method in the Macaronesian region. Harmful Algae
726 87, 101634.
- 727 Fleming, L.E., Baden, D.G., Bean, J.A., Weisman, R., Blythe, D.G., 1998. Marine Seafood Toxin
728 Diseases: Issues In Epidemiology & Community Outreach In: Reguera, B., Blanco, J.,
729 Fernandez, M.L., Wyatt, T. (Eds.), Harmful Algae. Xunta de Galicia and Intergovernmental
730 Oceanographic Commission of UNESCO, pp. 245–248.
- 731 Fraga, S., Rodríguez, F., Caillaud, A., Diogène, J., Raho, N., Zapata, M., 2011. *Gambierdiscus*
732 *excentricus* sp. nov. (Dinophyceae), a benthic toxic dinoflagellate from the Canary Islands

- 733 (NE Atlantic Ocean). Harmful Algae 11, 10-22.
- 734 Friedman, M.A., Fernandez, M., Backer, L.C., Dickey, R.W., Bernstein, J., Schrank, K., Kibler, S.,
735 Stephan, W., Gribble, M.O., Bienfang, P., Bowen, R.E., Degrasse, S., Quintana, H.A.F.,
736 Loeffler, C.R., Weisman, R., Blythe, D., Berdalet, E., Ayyar, R., Clarkson-Townsend, D.,
737 Swajian, K., Benner, R., Brewer, T., Fleming, L.E., 2017. An updated review of ciguatera fish
738 poisoning: Clinical, epidemiological, environmental, and public health management. Mar.
739 Drugs. 15(3): 72.
- 740 Fritz, L., Triemer, R.E., 1985. A Rapid simple technique utilizing calcofluor white M2R for the
741 visualization of dinoflagellate thecal plates. J. Phycol. 21, 662–664.
- 742 Garín-Muñoz, T., Montero-Martín, L.F., 2007. Tourism in the Balearic Islands: A dynamic model
743 for international demand using panel data. Tour. Manag. 28, 1224–1235.
- 744 Gatti, C.M.I., Lonati, D., Darius, H.T., Zancan, A., Roué, M., Schicchi, A., Locatelli, C.A., Chinain,
745 M., 2018. *Tectus niloticus* (Tegulidae, gastropod) as a novel vector of ciguatera poisoning:
746 Clinical characterization and follow-up of a mass poisoning event in Nuku Hiva Island
747 (French Polynesia). Toxins (Basel). 10.
- 748 Gómez, F., Qiu, D., Lopes, R.M., Lin, S., 2015. *Fukuyoa paulensis* gen. et sp. nov., a new genus for
749 the globular species of the dinoflagellate *Gambierdiscus* (Dinophyceae). PLoS One.
- 750 Gouveia, N.N., Vale, P., Gouveia, N., Delgado, J., 2010. Primeiro Registo da Ocorrência de
751 Episódios do Tipo Ciguatérico no Arquipélago da Madeira. Algas toxicas e biotoxinas nas
752 águas da Península Ibérica 152–157.
- 753 Guillard, R.R.L., 1973. Division rates, in: R., S.J. (Ed.), Handbook of Phycological Methods: Culture
754 Methods and Growth Measurements. Cambridge University Press, Cambridge, pp. 289-
755 312.
- 756 Guillard, R.R.L., Ryther, J.H., 1962. Studies of marine planktonic diatoms. I. *Cyclotella nana*

- 757 Hustedt, and *Detonula confervacea* Cleve. Can. J. Microbiol. 8, 229–239.
- 758 Hall, T.A., 1999. BIOEDIT: a user-friendly biological sequence alignment editor and analysis
759 program for Windows 95/98/ NT. Nucleic Acids Symp. Ser.
- 760 Hallegraeff, G.M., 2015. Transport of harmful marine microalgae via ship's ballast water:
761 Management and mitigation with special reference to the Arabian Gulf region. Aquat.
762 Ecosyst. Heal. Manag. 18, 290–298.
- 763 Herzberg, A., 1973. Toxicity of *Siganus luridus* (Rupell) on the Mediterranean Coast of Israel.
764 Aquaculture 2, 89–91.
- 765 Hidalgo, J., Liberona, J.L., Molgó, J., Jaimovich, E., 2002. Pacific ciguatoxin-1B effect over Na⁺
766 and K⁺ currents, inositol 1,4,5-triphosphate content and intracellular Ca²⁺ signals in
767 cultured rat myotubes. Br. J. Pharmacol. 137, 1055–62.
- 768 Holland, W.C., Litaker, R.W., Tomas, C.R., Kibler, S.R., Place, A.R., Davenport, E.D., Tester, P.A.,
769 2013. Differences in the toxicity of six *Gambierdiscus* (Dinophyceae) species measured
770 using an in vitro human erythrocyte lysis assay. Toxicon 65, 15-33.
- 771 Holmes, M.J., Lewis, R.J., Gillespie, N.C., 1990. Toxicity of Australian and French Polynesian
772 strains of *Gambierdiscus toxicus* (Dinophyceae) grown in culture: Characterization of a new
773 type of maitotoxin. Toxicon 28, 1159–1172.
- 774 Holmes, M.J., Lewis, R.J., 1994. Purification and characterisation of large and small maitotoxins
775 from cultured gambierdiscus toxicus. Nat. Toxins 2, 64–72.
- 776 Hoshaw, R. W., and Rosowski, J.R., 1973. Handbook of Phycological Methods: Culture Methods
777 and Growth Measurements., in: Stein, J.R. (Ed.). University Press, Cambridge, London and New
778 York, N.Y, p. 448.
- 779 Huelsenbeck, J.P., Ronquist, F., 2001. MRBAYES: Bayesian inference of phylogenetic trees.

- 780 Bioinformatics 17, 754–755.
- 781 Inserra, M.C., Israel, M.R., Caldwell, A., Castro, J., Deuis, J.R., Harrington, A.M., Keramidas, A.,
782 Garcia-Caraballo, S., Maddern, J., Erickson, A., Grundy, L., Rychkov, G.Y., Zimmermann, K.,
783 Lewis, R.J., Brierley, S.M., Vetter, I., 2017. Multiple sodium channel isoforms mediate the
784 pathological effects of Pacific ciguatoxin-1. *Sci. Rep.* 7:42810.
- 785 Jang, S.H., Jeong, H.J., Yoo, Y. Du, 2018. *Gambierdiscus jejuensis* sp. nov., an epiphytic
786 dinoflagellate from the waters of Jeju Island, Korea, effect of temperature on the growth,
787 and its global distribution. *Harmful Algae* 80, 149-157.
- 788 Kibler, S.R., Litaker, R.W., Holland, W.C., Vandersea, M.W., Tester, P.A., 2012. Growth of eight
789 *Gambierdiscus* (Dinophyceae) species: Effects of temperature, salinity and irradiance.
790 *Harmful Algae* 19, 1–14.
- 791 Kibler, S.R., Tester, P.A., Kunkel, K.E., Moore, S.K., Litaker, R.W., 2015. Effects of ocean warming
792 on growth and distribution of dinoflagellates associated with ciguatera fish poisoning in
793 the Caribbean. *Ecol. Modell.* 24, 194-210.
- 794 Kohler, S.T., Kohler, C.C., 1992. Dead bleached coral provides new surfaces for dinoflagellates
795 implicated in ciguatera fish poisonings. *Environ. Biol. Fishes* 35, 413–416.
- 796 Kohli, G. S., Neilan, B.A., Brown, M. V., Hoppenrath, M., Murray, S.A., 2014a. Cob gene
797 pyrosequencing enables characterization of benthic dinoflagellate diversity and
798 biogeography. *Environ. Microbiol.* 16(2):467-485.
- 799 Kohli, G. S., Papiol, G.G., Rhodes, L., Harwood, D.T., Selwood, A., Jerrett, A., Murray, S.A.,
800 Neilan, B.A., 2014b. A feeding study to probe the uptake of Maitotoxin by snapper
801 (*Pagrus auratus*). *Harmful Algae* 37, 125–132.
- 802 Larsson, M.E., Harwood, T.D., Lewis, R.J., Himaya, S.W.A., Doblin, M.A., 2019. Toxicological
803 characterization of *Fukuyoa paulensis* (Dinophyceae) from temperate Australia. *Phycol.*

- 804 Res. 67, 65–71.
- 805 Larsson, M.E., Laczka, O.F., Tim Harwood, D., Lewis, R.J., Himaya, S.W.A., Murray, S.A., Doblin,
806 M.A., 2018. Toxicology of *Gambierdiscus* spp. (dinophyceae) from tropical and temperate
807 Australian waters. *Mar. Drugs* 16, 1–19.
- 808 Laza-Martínez, A., David, H., Riobó, P., Miguel, I., Orive, E., 2016. Characterization of a Strain of
809 *Fukuyoa paulensis* (Dinophyceae) from the Western Mediterranean Sea. *J. Eukaryot.*
810 *Microbiol.* 63(4):481-497.
- 811 Lejeusne, C., Chevaldonné, P., Pergent-Martini, C., Boudouresque, C.F., Pérez, T., 2009. Climate
812 change effects on a miniature ocean: the highly diverse, highly impacted Mediterranean
813 Sea. *Trends Ecol. Evol.* 25(4):250-260.
- 814 Leung, P.T.Y., Yan, M., Lam, V.T.T., Yiu, S.K.F., Chen, C.Y., Murray, J.S., Harwood, D.T., Rhodes, L.,
815 Lam, P.K.S., Wai, T.C., 2018. Phylogeny, morphology and toxicity of benthic dinoflagellates
816 of the genus *Fukuyoa* (Goniodomataceae, Dinophyceae) from a subtropical reef ecosystem
817 in the South China Sea. *Harmful Algae* 74, 78–97.
- 818 Lewis, R.J., Holmes, M.J., 1993. Origin and transfer of toxins involved in ciguatera. *Comp.*
819 *Biochem. Physiol. Part C Comp.* 106, 615–628.
- 820 Litaker, R.W., Holland, W.C., Hardison, D.R., Pisapia, F., Hess, P., Kibler, S.R., Tester, P.A., 2017.
821 Ciguatoxicity of *Gambierdiscus* and *Fukuyoa* species from the Caribbean and Gulf of
822 Mexico. *PLoS One* 12, 1–19.
- 823 Litaker, R.W., Vandersea, M.W., Faust, M.A., Kibler, S.R., Chinain, M., Holmes, M.J., Holland,
824 W.C., Tester, P.A., 2009. Taxonomy of *Gambierdiscus* including four new species,
825 *Gambierdiscus caribaeus*, *Gambierdiscus carolinianus*, *Gambierdiscus carpenteri* and
826 *Gambierdiscus ruetzleri* (Gonyaulacales, Dinophyceae). *Phycologia* 48, 344–390.
- 827 Litaker, R.W., Vandersea, M.W., Faust, M.A., Kibler, S.R., Nau, A.W., Holland, W.C., Chinain, M.,

- 828 Holmes, M.J., Tester, P.A., 2010. Global distribution of ciguatera causing dinoflagellates in
829 the genus *Gambierdiscus*. *Toxicon*. 56(5):711-730.
- 830 Llewellyn, L.E., 2010. Revisiting the association between sea surface temperature and the
831 epidemiology of fish poisoning in the South Pacific: reassessing the link between ciguatera
832 and climate change. *Toxicon* 56, 691–697.
- 833 Longo, S., Sibat, M., Viallon, J., Darius, H.T., Hess, P., Chinain, M., 2019. Intraspecific variability
834 in the toxin production and toxin profiles of in vitro cultures of *Gambierdiscus polynesiensis*
835 (dinophyceae) from French polynesia. *Toxins (Basel)*. 11, 1–23.
- 836 Manger, R.L., Leja, L.S., Lee, S.Y., Hungerford, J.M., Kirkpatrick, M.A., Yasumoto, T., Wekell,
837 M.M., 2003. Detection of paralytic shellfish poison by rapid cell bioassay: Antagonism of
838 voltage-gated sodium channel active toxins in vitro. *J. AOAC Int.* 86(3): 540-543.
- 839 Manger, R.L., Leja, L.S., Lee, S.Y., Hungerford, J.M., Wekell, M.M., 1993. Tetrazolium-based cell
840 bioassay for neurotoxins active on voltage-sensitive sodium channels: Semiautomated
841 assay for saxitoxins, brevetoxins, and ciguatoxins. *Anal. Biochem.* 214(1):190-194.
- 842 Manger, R.L., Leja, L.S., Lee, S.Y., Jem M Hungerford, Yoshitsugi Hokama, R.W.D., Granade, H.R.,
843 Lewis, R., Takeshi Yasumoto, M.M.W., 1995. Detection of Sodium Channel Toxins: Directed
844 Cytotoxicity Assays of Purified Ciguatoxins, Brevetoxins, Saxitoxins, and Seafood Extracts.
845 *J. AOAC Int.* 78, 521–527.
- 846 Molgó, J., Shimahara, T., Legrand, A.M., 1993. Ciguatoxin, extracted from poisonous morays
847 eels, causes sodium-dependent calcium mobilization in NG108-15 neuroblastoma × glioma
848 hybrid cells. *Neurosci. Lett.* 158, 147–150.
- 849 Munday, R., 2014. Toxicology of Seafood Toxins: A Critical Review, in: Botana, L.M. (Ed.), *Seafood*
850 *and Freshwater Toxins: Pharmacology, Physiology, and Detection*. CRC Press, Boca Raton,
851 FL, pp. 197–290.

- 852 Munday, R., Murray, S., Rhodes, L., Larsson, M., Harwood, D., 2017. Ciguatoxins and Maitotoxins
853 in Extracts of Sixteen *Gambierdiscus* Isolates and One *Fukuyoa* Isolate from the South
854 Pacific and Their Toxicity to Mice by Intraperitoneal and Oral Administration. *Mar. Drugs*
855 15, 208.
- 856 Murray, J.S., Nishimura, T., Finch, S.C., Rhodes, L., Puddick, J., Harwood, D.T., Larsson, M.E.,
857 Doblin, M.A., Leung, P., Yan, M., Rise, F., Wilkins, A.L., Prinsep, M.R., 2020. The role of 44-
858 methylgambierone in ciguatera fish poisoning: acute toxicity, production by marine
859 microalgae and its potential as a biomarker for *Gambierdiscus* spp. *Harmful Algae* 97,
860 101853.
- 861 Murray, J.S., Selwood, A.I., Harwood, D.T., van Ginkel, R., Puddick, J., Rhodes, L., Rise, F., Wilkins,
862 A.L., 2019. 44-Methylgambierone, a new gambierone analogue isolated from
863 *Gambierdiscus australes*. *Tetrahedron Lett.* 60, 621–625.
- 864 Nascimento, S.M., Melo, G., Salgueiro, F., Diniz, B. dos S., Fraga, S., 2015. Morphology of
865 *Gambierdiscus excentricus* (Dinophyceae) with emphasis on sulcal plates. *Phycologia* 54(6).
- 866 Nicholson, G.M., Lewis, R.J., 2006. Ciguatoxins: Cyclic polyether modulators of voltage-gated ion
867 channel function. *Mar. Drugs* 4, 82–118.
- 868 Parsons, M.L., Preskitt, L.B., 2007. A survey of epiphytic dinoflagellates from the coastal waters
869 of the island of Hawai'i. *Harmful Algae* 6, 658–669.
- 870 Parsons, M.L., Settlemyer, C.J., Ballauer, J.M., 2011. An examination of the epiphytic nature of
871 *Gambierdiscus toxicus*, a dinoflagellate involved in ciguatera fish poisoning. *Harmful Algae*
872 10, 598–605.
- 873 Paz, B., Riobó, P., Franco, J.M., 2011. Preliminary study for rapid determination of phycotoxins
874 in microalgae whole cells using matrix-assisted laser desorption/ionization time-of-flight
875 mass spectrometry. *Rapid Commun. Mass Spectrom.* 25, 3627–3639.

- 876 Penna, A., Vila, M., Fraga, S., Giacobbe, M.G., Francesco, A., Riobó, P., Vernesi, C., 2005.
877 Characterization of *Ostreopsis* and *Coolia* (Dinophyceae) isolates in the western
878 Mediterranean Sea based on morphology, toxicity and internal transcribed spacer 5.8s
879 rDNA sequences. *J. Phycol.* 41(1):212-225.
- 880 Pisapia, F., Holland, W.C., Hardison, D.R., Litaker, R.W., Fraga, S., Nishimura, T., Adachi, M.,
881 Nguyen-Ngoc, L., Séchet, V., Amzil, Z., Herrenknecht, C., Hess, P., 2017. Toxicity screening
882 of 13 *Gambierdiscus* strains using neuro-2a and erythrocyte lysis bioassays. *Harmful Algae.*
883 63, 173-183.
- 884 Provasoli, L., 1968. Media and prospects of the cultivation of marine algae, in: *Culture and*
885 *Collection of Algae. Proceedings. Japanese Society of Plant Physiology, Hakone, Japan*, pp.
886 63–75.
- 887 Puigserver, M., Monerris, N., Moya, G., 2008. Estudi del fitoplàncton de les aigües costaneres de
888 les Illes Balears (2005-2006) en el marc de la implantació de la Directiva Marc Europea de
889 l'Aigua per a l'avaluació del seu estat ecològic. *Bolletí la Soc. d'Història Natural de les*
890 *Balears.* 51, 49–61.
- 891 Puigserver, M., Ramon, G., Moyà, G., Martínez-Taberne, A., 2002. Planktonic chlorophyll a and
892 eutrophication in two Mediterranean littoral systems (Mallorca Island, Spain)., in: Orive,
893 E., Elliott, M., de Jonge, V.N. (Eds.), *Nutrients and Eutrophication in Estuaries and Coastal*
894 *Waters. Developments in Hydrobiology. Springer Dordrecht*, pp. 493–504.
- 895 Raikhlin-Eisenkraft, B., Bentur, Y., 2002. Rabbitfish ("Aras"): An unusual source of Ciguatera
896 Poisoning. *Isr. Med. Assoc. J.* 4.
- 897 Raiklin-Eisenkraft, B., Finkelstein, Y., Spanier, E., 1988. Ciguatera-like Poisoning in the
898 Mediterranean. *Vet. Hum. Toxicol.* 30, 6.
- 899 Reverté, L., Toldrà, A., Andree, K.B., Fraga, S., de Falco, G., Campàs, M., Diogène, J., 2018.

- 900 Assessment of cytotoxicity in ten strains of *Gambierdiscus australes* from Macaronesian
901 Islands by neuro-2a cell-based assays. *J. Appl. Phycol.* 30, 2447–2461.
- 902 Reyes, J.G., Sánchez-Cárdenas, C., Acevedo-Castillo, W., Leyton, P., López-González, I., Felix, R.,
903 Gandini, M.A., Treviño, M.B., Treviño, C.L., 2014. Maitotoxin: An Enigmatic Toxic Molecule
904 with Useful Applications in the Biomedical Sciences, in: *Seafood and Freshwater Toxins*.
- 905 Rhodes, L., Giménez Papiol, G., Smith, K., Harwood, T., 2014a. *Gambierdiscus* cf. *yasumotoi*
906 (Dinophyceae) isolated from New Zealand’s sub-tropical northern coastal waters. *New*
907 *Zeal. J. Mar. Freshw. Res.* 48, 303–310.
- 908 Rhodes, L., Harwood, T., Smith, K., Argyle, P., Munday, R., 2014b. Production of ciguatera and
909 maitotoxin by strains of *Gambierdiscus australes*, *G. pacificus* and *G. polynesiensis*
910 (Dinophyceae) isolated from Rarotonga, Cook Islands. *Harmful Algae* 39, 185-190.
- 911 Rhodes, L., Smith, K.F., Murray, S., Harwood, D.T., Trnski, T., Munday, R., 2017. The epiphytic
912 genus *Gambierdiscus* (Dinophyceae) in the Kermadec Islands and Zealandia regions of the
913 southwestern Pacific and the associated risk of ciguatera fish poisoning. *Mar. Drugs*.
914 15(7):219.
- 915 Rodríguez, F., Fraga, S., Ramilo, I., Rial, P., Figueroa, R.I., Riobó, P., Bravo, I., 2017. “Canary Islands
916 (NE Atlantic) as a biodiversity ‘hotspot’ of *Gambierdiscus*: Implications for future trends of
917 ciguatera in the area.” *Harmful Algae* 67, 131–143.
- 918 Roeder, K., Eler, K., Kibler, S., Tester, P., Van The, H., Nguyen-Ngoc, L., Gerdt, G., Luckas, B.,
919 2010. Characteristic profiles of Ciguatera toxins in different strains of *Gambierdiscus* spp.
920 *Toxicon* 56, 731–738.
- 921 Rongo, T., van Woesik, R., 2013. The effects of natural disturbances, reef state, and herbivorous
922 fish densities on ciguatera poisoning in Rarotonga, southern Cook Islands. *Toxicon* 64, 87–
923 95.

- 924 Rossignoli, A.E., Tudó, A., Bravo, I., Díaz, P.A., Diogène, J., Riobó, P., 2020. Toxicity
925 characterisation of *Gambierdiscus* species from the Canary Islands. *Toxins* (Basel). 12, 1–
926 15.
- 927 Roué, M., Darius, H.T., Picot, S., Ung, A., Viallon, J., Gaertner-Mazouni, N., Sibat, M., Amzil, Z.,
928 Chinain, M., 2016. Evidence of the bioaccumulation of ciguatoxins in giant clams (*Tridacna*
929 *maxima*) exposed to *Gambierdiscus* spp. cells. *Harmful Algae* 57, 78-87.
- 930 Rozewicki, J., Li, S., Amada, K.M., Standley, D.M., Katoh, K., 2019. MAFFT-DASH: integrated
931 protein sequence and structural alignment. *Nucleic Acids Res.* 47, W5–W10.
- 932 Saburova, M., Polikarpov, I., Al-Yamani, F., 2013. New records of the genus *Gambierdiscus* in
933 marginal seas of the Indian Ocean. *Mar. Biodivers. Rec.* 6.
- 934 Sakai, A.K., Allendorf, F.W., Holt, J.S., Lodge, D.M., Molofsky, J., With, K.A., Baughman, S., Cabin,
935 R.J., Cohen, J.E., Ellstrand, N.C., McCauley, D.E., O’Neil, P., Parker, I.M., Thompson, J.N.,
936 Weller, S.G., 2001. The population biology of invasive species. *Annu. Rev. Ecol. Syst.* 32,
937 305–332.
- 938 Satake, M., Isidbashi, Y., Legrand, A.M., Yasumoto, T., 1996. Isolation and structure of
939 ciguatoxin-4a, a new ciguatoxin precursor, from cultures of dinoflagellate *Gambierdiscus*
940 *toxicus* and parrotfish *Scarus gibbus*. *Biosci. Biotechnol. Biochem.* 60, 2103–2105.
- 941 Schneider, C.A., Rasband, W.S., Eliceiri, K.W., 2012. NIH Image to ImageJ: 25 years of image
942 analysis. *Nat. Methods.*
- 943 Singh, A., Hildebrand, M.E., Garcia, E., Snutch, T.P., 2010. The transient receptor potential
944 channel antagonist SKF96365 is a potent blocker of low-voltage-activated T-type calcium
945 channels. *Br. J. Pharmacol.* 160, 1464–1475.
- 946 Skinner, M.P., Brewer, T.D., Johnstone, R., Fleming, L.E., Lewis, R.J., 2011. Ciguatera fish
947 poisoning in the pacific islands (1998 to 2008). *PLoS Negl. Trop. Dis.* 5, 1–7.

- 948 Skinner, M.P., Lewis, R.J., Morton, S., 2013. Ecology of the ciguatera causing dinoflagellates from
949 the Northern Great Barrier Reef: Changes in community distribution and coastal
950 eutrophication. *Mar. Pollut. Bull.* 77, 210–219.
- 951 Spanier, E., Finkelstein, Y., Raikhlin-Eisenkraft, B., 1989. Toxicity of the saupe, *Sarpa salpa*
952 (Linnaeus, 1758), on the Mediterranean coast of Israel. *J. Fish Biol.* 34, 635–636.
- 953 Stamatakis, A., 2014. RAxML version 8: A tool for phylogenetic analysis and post-analysis of large
954 phylogenies. *Bioinformatics* 30, 1312–1313.
- 955 Strachan, L.C., Lewis, R.J., Nicholson, G.M., 1999. Differential Actions of Pacific Ciguatoxin-1 on
956 Sodium Channel Subtypes in Mammalian Sensory Neurons. *J. Pharmacol. Exp. Ther.*
- 957 Tester, P.A., Litaker, R.W., Berdalet, E., 2020. Climate change and harmful benthic microalgae.
958 *Harmful Algae* 91, 101655.
- 959 Tester, P.A., Vandersea, M.W., Buckel, C.A., Kibler, S.R., Holland, W.C., Davenport, E.D., Clark,
960 R.D., Edwards, K.F., Taylor, J.C., Pluym, J.L.V., Hickerson, E.L., Litaker, R.W., 2013.
961 *Gambierdiscus* (Dinophyceae) species diversity in the flower garden banks national marine
962 sanctuary, Northern Gulf of Mexico, USA. *Harmful Algae*. 29, 1-9.
- 963 Toldrà, A., Andree, K.B., Fernández-Tejedor, M., Diogène, J., Campàs, M., 2018. Dual quantitative
964 PCR assay for identification and enumeration of *Karlodinium veneficum* and *Karlodinium*
965 *armiger* combined with a simple and rapid DNA extraction method. *J. Appl. Phycol.* 30.
- 966 Tudó, À.; Toldrà, A.; Andree, K. B.; Rey, M.; Fernández-Tejedor M.; Campàs, M.; Diogène, J.,
967 2018. First report of *Gambierdiscus* in the Western Mediterranean Sea (Balearic Islands).
968 *Harmful Algae News* 59, 22–23.
- 969 Turquet, J., Jean-Pascal, Q., Ten-Hage, L., Dahalani, Y., Wendling, B., 2001. Example of a
970 *Gambierdiscus toxicus* flare-up following the 1998 coral bleaching event in Mayotte Island
971 (Comoros, South-west Indian Ocean), in: Hallegraeff, G.M., Blackburn, S.I., Bolch, C.J.,

- 972 Lewis, R.J. (Eds.), Harmful Algal Blooms 2001, Intergovernmental Oceanographic
973 Commission of UNESCO. pp. 50–53.
- 974 Vila, M., Garcés, E., Masó, M., 2001a. Potentially toxic epiphytic dinoflagellate assemblages on
975 macroalgae in the NW Mediterranean. *Aquat. Microb. Ecol.* 26(1):51-60.
- 976 Vila, M., Camp, J., Garcés, E., Masó, M., Delgado, M. 2001b. High resolution spatio-temporal
977 detection of potentially harmful dinoflagellates in confined waters of the NW
978 Mediterranean. *J. Plankton Res.* 23(5):497-514.
- 979 Vila, M., Garcés, E., Masó, M., Camp, J., 2001c. Is the distribution of the toxic dinoflagellate
980 *Alexandrium catenella* expanding along the NW Mediterranean coast? *Mar. Ecol. Prog. Ser.*
981 222, 73–83.
- 982 Villareal, T.A., Hanson, S., Qualia, S., Jester, E.L.E., Granade, H.R., Dickey, R.W., 2007. Petroleum
983 production platforms as sites for the expansion of ciguatera in the northwestern Gulf of
984 Mexico. *Harmful Algae* 6, 253–259.
- 985 Whithers, N., 1981. Toxin production, nutrition and distribution *Gambierdiscus toxicus* (Hawaiian
986 strain), in: Fourth International Coral Reef Symposium. pp. 449-451.
- 987 Wood, M.A., Everroad, C.R., Wingard, M.L., 2005. Measuring growth rates in microalgal cultures,
988 in: Andersen, R.A. (Ed.), *Algal Culturing Techniques*. Elsevier Academic Press, pp. 269–286.
- 989 Xu, Y., Richlen, M.L., Liefer, J.D., Robertson, A., Kulis, D., Smith, T.B., Parsons, M.L., Anderson,
990 D.M., 2016. Influence of Environmental Variables on *Gambierdiscus* spp. (Dinophyceae)
991 Growth and Distribution. *PLoS One.* 11(4).
- 992 Yasumoto, T., Bagnis, R., Thevenin, S., Garcon, M., 1977. A Survey of Comparative Toxicity in the
993 Food Chain of Ciguatera. *Nippon SUI SAN GAKKAISHI*.
- 994 Yasumoto, T., Hashimoto, Y., Bagnis, R., Randall, J.E., Banner, A.H., 1971. Toxicity of the

- 995 Surgeonfishes. Nippon SUISAN GAKKAISHI.
- 996 Yoshimatsu, T., Yamaguchi, H., Iwamoto, H., Nishimura, T., Adachi, M., 2014. Effects of
997 temperature, salinity and their interaction on growth of Japanese *Gambierdiscus* spp.
998 (Dinophyceae). Harmful Algae 35, 29–37.
- 999 Zenetos, A., Gofas, S., Morri, C., Rosso, A., Violanti, D., Garcia Raso, J.E., Cinar, M.E., Almogi-
1000 Labin, A., Ates, A.S., Azzurro, E., Ballesteros, E., Bianchi, C.N., Bilecenoglu, M., Gambi, M.C.,
1001 Giangrande, A., Gravili, C., Hyams-Kaphzan, O., Karachle, P.K., Katsanevakis, S., Lipej, L.,
1002 Mastrototaro, F., Mineur, F., Pancucci-Papadopoulou, M.A., Ramos Espla, A., Salas, C., San
1003 Martin, G., Sfriso, A., Streftaris, N., Verlaque, M., 2012. Alien species in the Mediterranean
1004 Sea by 2012. A contribution to the application of European Union’s Marine Strategy
1005 Framework Directive (MSFD). Part 2. Introduction trends and pathways. Mediterr. Mar. Sci.
1006 13, 328.
- 1007
- 1008
- 1009
- 1010
- 1011
- 1012
- 1013
- 1014
- 1015
- 1016
- 1017

1018 **Table. 1.** Morphometric comparison of *G. australes* and *F. paulensis* strains of this study with
1019 published measurements for those species. Average and standard deviation of depth (D), length
1020 (L), width (W), ratio of depth and width (D:W), ratio of length and width (L:W); length of apical
1021 pore plate (Po) and surrounding pore numbers (No) and diameter (ϕ), size of 2'''' plate. Data are
1022 expressed as the arithmetic mean, standard deviation (\pm SD) and number of measured cells (n).

		<i>G. australes</i> Chinain et al. 1999	<i>G. australes</i> Litaker et al., 2009	<i>G. australes</i> Rhodes et al., 2014b	<i>G. australes</i> Rhodes et al., 2014b	<i>G. australes</i> Bravo et al. 2019	<i>G. australes</i> This study	<i>F. paulensis</i> Rhodes et al., 2014a (<i>G. yasumotoi</i>)	<i>F. paulensis</i> Gómez et al. 2015	<i>F. paulensis</i> Laza-Martínez et al. 2016	<i>F. paulensis</i> This study
Isolates		RAV-92	RAV-92/NOAA24	CADW149	CAWD216			CAWD210	VGO1185	Dn35EHU	
Cell size	L μm	-	38.7 \pm 3.8 (33.4 - 47.3)	32.0 (26.0 - 39.0)	39.0 (32.5 - 45.5)		-	59.8 \pm 7.5 (54.3- 67.3) (n=20)	56.0 \pm 3.0 (51- 62)	48.9 \pm 10.9 (35- 76) (n=100)	46.6 \pm 8.7 (32.0- 64.1) (n=21)
	D μm	86.0 \pm 5.1 (76.0 - 93.0)	72.5 \pm 3.8 (63.8 - 77.4)	44.5 (32.5 - 52.0)	58.5 (45.5 - 65.0)	81 \pm 6.3 (68-95)	75.7 \pm 6.0 (60.9- 92.3) (n= 112)	54.8 \pm 5.7 (49.1- 60.5) (n=20)	50.0 \pm 3.0 (45- 56)	40.8 \pm 8.2 (31- 67) (n=123)	40.5 \pm 4.8 (36.4 - 51.1) (n=14)
	W μm	77.0 \pm 3.7 (65.0 - 84.0)	63.4 \pm 5 (55.2-73.8)	38.5 (32.5-52.0)	48.0 (40.0-52.0)	78 \pm 7.5 (60-95)	78.7 \pm 6.6 (54.7 - 90.8) (n= 112)	42.5 \pm 4.1 (38.4- 46.6) (n=20)	45.0 \pm 2.0 (41- 48)	30.5 \pm 6.6 (24- 38) (n=60)	41.1 \pm 11.9 (11.9 - 41.1) (n=21)
	L:W		0.61	0.83	0.81			1.41	n.d	1.28 (n=10)	1.44 \pm 0.21 (1.1 - 1.7) (n=21)
	D:W	1.12	1.14	1.16	1.22		1.02 \pm 0.09 (0.8 - 1.2) (n= 112)	1.29	1.2	1.29 (n=48)	
Po plate	L μm	7.1 \pm 0.8 (6.3-8.6)					7.2 \pm 0.7 (6.2 - 8.4) (n=14)	9.9 (Laza- Martínez et al., 2016)	10-12	7.6	8.2 \pm 1.4 (5.9 - 11.4) (n=13)
	W μm	6.1 \pm 0.4 (5.7-6.8)					5.7 \pm 0.7 (4.8 - 7.8) (n=14)	4.6 (Laza- Martínez et al., 2016)	6-7	4.1	3.0 \pm 0.7 (2.2-4.4) (n=13)
	L:W	1.18					1.23 \pm 0.1 (1.0- 1.5) (n=14)		n.d		3.1 \pm 0.9 (1.8-4.4) (n=13)
	Number pores	31 \pm 4.1					29 \pm 1.6 (27- (n=14)	33)	23-39	29-39	35.3 \pm 1.6 (32- (n=13)
	Diameter pores μm	0.45 \pm 0.03					0.39 \pm 0.09 (0.2- 0.6) (n=197)			0.35 (n=150)	0.31 \pm 0.08 (0.16 - 0.51) (n=52)
2'''' antapical	L 2'''' μm	54 \pm 3.1					41.5 \pm 4.8 (33.5 - 48.8) (n=14)		33-39		45.3 \pm 2.9 (41.7- 48.8)
	W 2'''' μm	27 \pm 2.7					21.6 \pm 3.3 (17.7 - 29.6) (n=14)		19-23		23.7 \pm 3.4 (19.4 - 29.6)
	L:W 2''''	2.10					1.95 \pm 0.28 (1.6 - 2.5) (n=14)				1.95 \pm 0.32 (1.7 - 2.5) (n=7)

1035 **Table. 2.** Growth parameters of *G. australes* (n=3) and *F. paulensis* (n=1) from the Balearic
 1036 Islands. Averages of the three replicates of: Max. conc.= maximum cell yield (cells · mL⁻¹); r =
 1037 growth rate (div. · day⁻¹) ± standard deviation (SD), the period when r was calculated (days) is
 1038 showed in brackets; K= doublings per day (doublings · day⁻¹) ± SD; Td = doubling time (days⁻¹) ±
 1039 SD.

	IRTA-SMM-17-162 <i>G. australes</i>	IRTA-SMM-17-189 <i>G. australes</i>	IRTA-SMM-17-271 <i>G. australes</i>	IRTA-SMM-17-209 <i>F. paulensis</i>
Max. conc.	2288	1451	1244	1004
r	0.12 ± 0.04 (13-20)	0.15 ± 0.04 (13-20)	0.16 ± 0.04 (12-19)	0.24 ± 0.06 (7-14)
K (Eq. 1)	0.17 ± 0.05	0.21 ± 0.06	0.24 ± 0.06	0.34 ± 0.09
Td (Eq. 2)	6.25 ± 1.80	5.85 ± 1.07	4.36 ± 1.01	1.30 ± 1.64

1040

1041

1042

1043

1044

1045

1046

1047

1048

1049

1050

1051

1052

1053 **Table 3.** Evaluation of the presence of CTX-like and MTX-like toxicity by neuro-2a CBA. Species,
 1054 code of strain, origin, cell concentration of cultures at harvesting time ($\text{cell} \cdot \text{mL}^{-1}$), values of CTX-
 1055 like toxicity expressed in femtograms (fg) of CTX1B equiv. $\cdot \text{cell}^{-1} \pm \text{SD}$. n.s.: nonspecific toxicity;
 1056 +: recovery of the cell viability in the presence of SKF96365; -: non-recovery of the cell
 1057 viability in the presence of SKF96365; NT: not tested.

Species	Code	Island	Cell abundance ($\text{cells} \cdot \text{mL}^{-1}$)	CTX-like toxicity (fg CTX1B equiv. $\cdot \text{cell}^{-1}$)	MTX-like Toxicity
<i>G. australes</i>	IRTA-SMM-17-153	Majorca	1750	1.38 \pm 0.66	+
<i>G. australes</i>	IRTA-SMM-17-238	Majorca	1632	3.52 \pm 0.18	NT
<i>G. australes</i>	IRTA-SMM-17-180	Minorca	613	5.25 \pm 0.59	NT
<i>G. australes</i>	IRTA-SMM-17-218	Majorca	1686	9.47 \pm 3.18	+
<i>G. australes</i>	IRTA-SMM-17-216 ^a	Majorca	1476	13.14 \pm 4.50	+
<i>G. australes</i>	IRTA-SMM-17-254	Majorca	1273	13.16 \pm 1.34	+
<i>G. australes</i>	IRTA-SMM-17-253	Majorca	1935	13.45 \pm 0.97	+
<i>G. australes</i>	IRTA-SMM-17-181	Minorca	1464	13.50 \pm 0.80	+
<i>G. australes</i>	IRTA-SMM-17-178	Minorca	2040	14.52 \pm 4.31	-
<i>G. australes</i>	IRTA-SMM-17-223	Majorca	1183	14.93 \pm 4.69	+
<i>G. australes</i>	IRTA-SMM-17-155	Minorca	332	17.33 \pm 1.60	-
<i>G. australes</i>	IRTA-SMM-17-173	Majorca	2087	21.89 \pm 9.20	+
<i>G. australes</i>	IRTA-SMM-17-244	Majorca	924	34.33 \pm 4.18	+
<i>G. australes</i>	IRTA-SMM-17-256	Majorca	1004	39.17 \pm 16.44	NT
<i>G. australes</i>	IRTA-SMM-17-175	Minorca	1498	62.00 \pm 0.66	-
<i>G. australes</i>	IRTA-SMM-17-164	Minorca	1022	72.60 \pm 43.20	+
<i>G. australes</i>	IRTA-SMM-17-214	Majorca	1694	76.67 \pm 29.86	+
<i>G. australes</i>	IRTA-SMM-17-189	Minorca	869	83.39 \pm 12.14	NT
<i>G. australes</i>	IRTA-SMM-17-162	Minorca	1390	105.67 \pm 18.27	NT
<i>G. australes</i>	IRTA-SMM-17-271	Minorca	843	172.63 \pm 5.57	+
<i>G. australes</i>	IRTA-SMM-17-168 ^a	Majorca	2274	381.83 \pm 91.84	NT
<i>F. paulensis</i>	IRTA-SMM-17-209 ^a	Minorca	782	16.30 \pm 1.67	NT
<i>F. paulensis</i>	IRTA-SMM-17-211	Minorca	3250	7.96 \pm 0.14	NT
<i>F. paulensis</i>	IRTA-SMM-17-198 ^a	Majorca	4825	n.s	NT
<i>F. paulensis</i>	IRTA-SMM-17-206	Majorca	2053	n.s	-
<i>F. paulensis</i>	IRTA-SMM-17-220	Minorca	2128	n.s	-
<i>F. paulensis</i>	IRTA-SMM-17-221	Minorca	6636	n.s	NT

1058 ^a response curves of CTX-like evaluation of these strains are shown in Fig. 5 and 6.

1059

1060

1061

1062