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1           **Cyclodextrin polymers as passive sampling materials for lipophilic**  
2           **marine toxins in *Prorocentrum lima* cultures and a *Dinophysis sacculus***  
3           **bloom in the NW Mediterranean Sea**

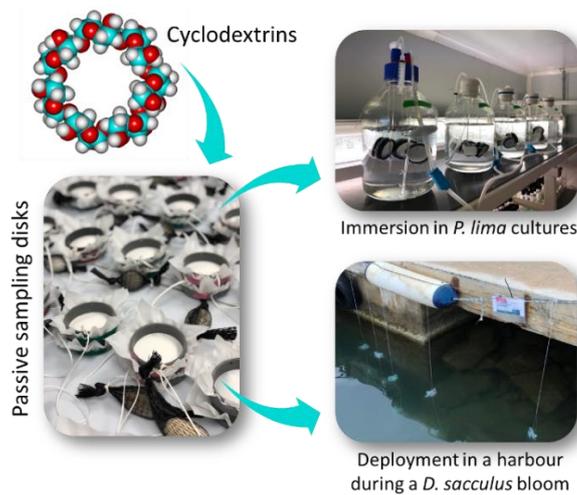
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10  
11       **Graphical abstract**



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14

15 **Abstract**

16 Cyclodextrins, cyclic oligomers that form a conical structure with an internal cavity, are proposed  
17 as new and sustainable materials for passive sampling of lipophilic marine toxins. Two  
18 applicability scenarios have been tested. First, disks containing  $\beta$ -cyclodextrin-hexamethylene  
19 diisocyanate ( $\beta$ -CD-HDI) and  $\beta$ -cyclodextrin-epichlorohydrin ( $\beta$ -CD-EPI) polymers were  
20 immersed in *Prorocentrum lima* cultures for different days (2, 12 and 40). LC-MS/MS analysis  
21 showed capture of free okadaic acid (OA) and dinophysistoxin-1 (DTX1) by cyclodextrins at  
22 contents that increased with immersion time. Cyclodextrins resulted more efficient in capturing  
23 DTX1 than OA. In a second experiment, disks containing  $\beta$ -CD-HDI,  $\beta$ -CD-EPI,  $\gamma$ -CD-HDI and  $\gamma$ -CD-  
24 EPI were deployed in harbor waters of El Masnou (NW Mediterranean Sea) during a *Dinophysis*  
25 *sacculus* bloom in February 2020. Free OA and pectenotoxin-2 (PTX2) were captured by  
26 cyclodextrins. Toxin contents were higher at sampling points and sampling weeks with higher  
27 *D. sacculus* cell abundance. In this case, PTX2 capture with cyclodextrins was more efficient than  
28 OA capture. Therefore, cyclodextrins have provided information regarding the toxin profile of a  
29 *P. lima* strain and the spatial and temporal dynamics of a *D. sacculus* bloom, proven efficient as  
30 passive sampling materials for environmental monitoring.

31 **Keywords**

32 Cyclodextrin, okadaic acid (OA), dinophysistoxin-1 (DTX1), pectenotoxin-2 (PTX2), *Prorocentrum*  
33 *lima*, *Dinophysis sacculus*.

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## 39 **1. Introduction**

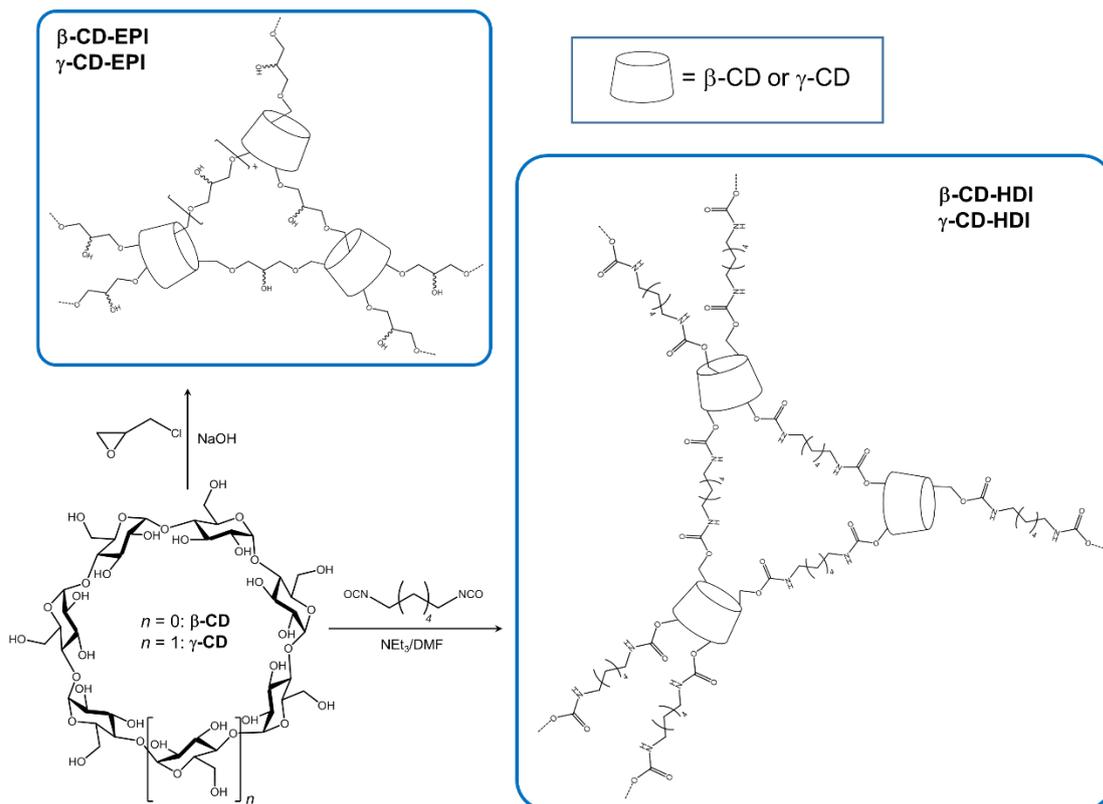
40 Harmful algal blooms (HABs) are increasing in geographical expansion, frequency and severity,  
41 some of the possible reasons being ocean warming, eutrophication and globalization (Wells et  
42 al., 2020). HABs represent a threat for food safety and consumers, since toxins produced by toxic  
43 microalgae are accumulated by shellfish and fish. Regarding shellfish safety, monitoring  
44 programs involve sampling of shellfish for analysis of marine toxins as well sampling of seawater  
45 for phytoplankton identification and counting. Early warning of HABs and shellfish contamination  
46 would be a useful approach to facilitate management and protect human health.

47 Solid-phase adsorption toxin tracking (SPATT) was conceived by MacKenzie in 2004 (MacKenzie  
48 et al., 2004). The technique involves the passive adsorption of toxins on porous synthetic resins  
49 for the subsequent extraction and analysis. These resins are encased into different types of  
50 frames and are deployed into the water column, where they can adsorb the toxins released by  
51 toxic microalgae. Passive samplers have been proposed as an early warning tool to forecast  
52 shellfish contamination or at least as complementary tool in monitoring programs. The main  
53 advantages of passive samplers are simplicity, low cost, low matrix effects when analyzing the  
54 resins, spatially and temporally integrated responses, and accumulative capacity. However, some  
55 issues still need to be resolved, such as optimal deployment times, saturation limits, lack of  
56 calibration and standardization, and the insufficient knowledge about the correlation with toxin  
57 contents in shellfish. Additionally, a major limitation is that they allow the monitoring of  
58 dissolved toxins only. Nevertheless, passive samplers already have an undeniable application in  
59 environmental monitoring and research, since they can provide information on HABs such as  
60 geographical and temporal distribution, environmental persistence and toxin dynamics.

61 Diaion® HP-20 has been the adsorbent substrate most commonly used in passive samplers (Roué  
62 et al., 2018) both in microalgae cultures (Fux et al., 2008; Li et al., 2011; Kudela, 2017) and field  
63 studies (MacKenzie et al., 2004; Zendong et al., 2015; Kudela, 2017). Unlike this resin, which

64 adsorbs toxins, cyclodextrins (CDs) can capture organic compounds by supramolecular  
65 interactions, which may result in different affinity, kinetics and saturation behaviors.  
66 Cyclodextrins are cyclic  $\alpha$ -1 $\rightarrow$ 4-linked glucose oligomers that form a conical structure with an  
67 essentially hydrophobic internal cavity filled with disordered water molecules and two external  
68 hydrophilic rims decorated with hydroxyl groups. The number of glucose units in the most  
69 common cyclodextrins, 6 in  $\alpha$ -CD, 7 in  $\beta$ -CD and 8 in  $\gamma$ -CD, dictates the size of the cavity, which  
70 allows the inclusion of a variety of organic molecules of appropriate size, shape and polarity  
71 (Villalonga et al., 2007). They have been exploited in different fields, such as in targeted therapy  
72 as drug carriers (Ramirez et al., 2006a, 2006b, 2007), and in biosensors for electrode surface  
73 modification and signal amplification (Ortiz et al., 2011a, 2011b, 2011c, 2011d, 2012, 2014; Wajs  
74 et al., 2014, 2016). Nevertheless, until now they had never been used for marine toxin tracking.

75 In this work, several insoluble cyclodextrin polymers ( $\beta$ -cyclodextrin-hexamethylene  
76 diisocyanate ( $\beta$ -CD-HDI),  $\beta$ -cyclodextrin-epichlorohydrin ( $\beta$ -CD-EPI),  $\gamma$ -cyclodextrin-  
77 hexamethylene diisocyanate ( $\gamma$ -CD-HDI) and  $\gamma$ -cyclodextrin-epichlorohydrin ( $\gamma$ -CD-EPI)) (Fig. 1)  
78 have been immersed in *Prorocentrum lima* cultures and deployed in harbor waters during a  
79 *Dinophysis sacculus* bloom for their evaluation as new and sustainable passive sampling  
80 materials. The commercial Diaion<sup>®</sup> HP-20 has been used as a control. Results have been useful  
81 to obtain information about the toxin profile of the *P. lima* strain and to elucidate the toxin  
82 production, development and dynamics of the *D. sacculus* HAB.



83

84 **Figure 1.** Schematic representations of  $\beta$ -CD-HDI,  $\beta$ -CD-EPI,  $\gamma$ -CD-HDI and  $\gamma$ -CD-EPI and their syntheses  
 85 from native  $\beta$ -CD and  $\gamma$ -CD.

## 86 2. Materials and methods

### 87 2.1. Reagents and materials

88  $\beta$ -CD-HDI and  $\gamma$ -CD-HDI were synthesized by crosslinking the native dried CDs with  
 89 hexamethylene diisocyanate (1:7 and 1:8 molar ratio, respectively) in dimethylformamide  
 90 containing triethylamine (Mohamed et al., 2011).  $\beta$ -CD-EPI and  $\gamma$ -CD-EPI were prepared by  
 91 reaction of the native CDs with epichlorohydrin (1:14 and 1:16 molar ratio, respectively) in NaOH  
 92 (Crini et al., 1998) (Fig. 1). The products were purified by Soxhlet extraction with EtOH and water.  
 93 Diaion® HP-20 Supelco resin was obtained from VidraFoc (Barcelona, Spain). Certified reference  
 94 material of okadaic acid (OA) ( $15.56 \mu\text{g mL}^{-1}$  in MeOH) was obtained from CIFGA (Lugo, Spain).  
 95 Dinophysistoxin-1 (DTX1) ( $8.52 \mu\text{g mL}^{-1}$  in MeOH) and pectenotoxin-2 (PTX2) ( $4.40 \mu\text{g mL}^{-1}$  in  
 96 MeOH) were obtained from the National Research Council of Canada (NRC, Halifax, Canada).

97 Passive sampling disks were constructed by placing 1 g (cultures) or 10 g (harbor) of  $\beta$ -CD-HDI,  
98  $\beta$ -CD-EPI,  $\gamma$ -CD-HDI,  $\gamma$ -CD-EPI or Diaion® HP-20 between two layers of 1  $\mu$ m nylon mesh (Sefar  
99 Maissa S.A.U., Cardedeu, Barcelona, Spain), clipped between two cylindrical PVC rings (4-cm  
100 diameter for immersion in cultures and 7-cm diameter for deployment in a harbor) (Fig. 1SA).  
101 The passive sampling disks to be deployed in the harbor were provided with a counterweight to  
102 ensure stability. Cyclodextrins and resin were activated by soaking the disks in MeOH for 15 min  
103 and rinsing them with milli-Q water.

## 104 **2.2. Immersion of cyclodextrins in *Prorocentrum lima* cultures**

105 Clonal cultures of *P. lima* strain IRTA-SMM-17-47 (GenBank accession number: MW328564) from  
106 IRTA collection were grown in modified ES medium (Provasoli, 1968), first in Nunclon™ cell  
107 culture polystyrene flasks (Thermo Fisher Scientific) and afterwards in glass bottles. Modified ES  
108 medium was prepared with sterile aged seawater obtained from L'Ametlla de Mar (Spain),  
109 Mediterranean Sea (40.8465° N; 0.77243° E) at 10 m depth, which was passed through an  
110 activated carbon-PTFE membrane filter (Thermo Fisher Scientific) and a 0.22- $\mu$ m cellulose  
111 acetate filter (Merck KGaA, Germany). The salinity was adjusted to 36 with milli-Q water. Cultures  
112 were maintained at  $24 \pm 0.5$  °C under a light intensity of  $110 \mu\text{mol photons m}^{-2} \text{s}^{-1}$  with a 12:12 h  
113 light:dark regime. Passive sampling disks containing of  $\beta$ -CD-HDI,  $\beta$ -CD-EPI or Diaion® HP-20  
114 were immersed into *P. lima* cultures (3 disks per glass bottle) at day 0, and collected at day 2, day  
115 12 and day 40 (Fig. 1SB) (at the time of the experiment,  $\gamma$ -CD-HDI and  $\gamma$ -CD-EPI were not  
116 available). A culture with no passive sampling disks was used as a control to evaluate if their  
117 presence had any effect on the culture growth. Two aliquots of each culture were taken every  
118 few days, fixed with 3% Lugol's iodine, and cells were counted in duplicate using a Kolkwitz  
119 chamber (Hydro-Bios, Altenholz, Germany) under an inverted light microscope (Leica DMIL,  
120 Spain). Cultures (~4 L) were harvested at day 40 through vacuum filtration using a 5- $\mu$ m nylon  
121 mesh (Sefar Maissa S.A.U., Spain).

### 122 **2.3. Deployment of cyclodextrins in a harbor during a *Dinophysis sacculus* bloom**

123 Two consecutive deployments lasting 7 days (from 14/02/2020 to 21/02/2020, and from  
124 21/02/2020 to 28/02/2020) were performed at 5 sampling points of El Masnou harbor (NW  
125 Mediterranean Sea) during a *D. sacculus* bloom (Fig. 2S). Passive sampling disks containing of  $\beta$ -  
126 CD-HDI,  $\beta$ -CD-EPI,  $\gamma$ -CD-HDI,  $\gamma$ -CD-EPI or Diaion® HP-20 (in duplicate) were deployed at 1.2 m  
127 depth and left for 1 week (Fig. 1SC). Phytoplankton cells were counted under an inverted light  
128 microscope (Leica DMIL, Spain), following the Utermöhl method (Utermöhl, 1931).

### 129 **2.4. Toxin extraction**

130 The passive sampling disks were soaked in milli-Q water for 30 min. Afterwards, the embroidery  
131 hoop was opened, and the  $\beta$ -CD-HDI,  $\beta$ -CD-EPI,  $\gamma$ -CD-HDI,  $\gamma$ -CD-EPI or Diaion® HP-20 were  
132 transferred to beakers and incubated with MeOH (40 mL when using 1 g (cultures) and 80 mL  
133 when using 10 g (harbor)) for 2 h. Cyclodextrins and resin were then transferred to low frequency  
134 polyvinyl chloride (LPVC) plastic filtration columns containing 1  $\mu$ m nylon mesh filters and frits,  
135 vacuum was applied with a Vac-Elut SPE vacuum manifold (Varian, Harbor City, CA, USA), and the  
136 MeOH was collected. Rinsing was performed with additional MeOH (20-30 mL approx.), which  
137 was also collected. The total volume of eluate was evaporated to dryness in a Syncore Buchi  
138 (Flawil, Switzerland) and redissolved in 0.5 (when using 1 g) or 4 mL (when using 10 g) of MeOH.  
139 *Prorocentrum lima* culture media (0.5 L) were filtered through Empore™ C18 SPE Disks (Supelco,  
140 Sigma-Aldrich, Tres Cantos, Madrid, Spain). Disks were first conditioned with 10 mL of MeOH and  
141 10 mL of milli-Q water. Then, culture media were loaded, vacuum was applied, and the collected  
142 media were discarded. Samples were then eluted with 20 mL of MeOH. *Prorocentrum lima*  
143 cultures filters were sonicated 3 times in 150 mL of MeOH for 30 min. The three extracts were  
144 joined and centrifuged at 3,000 rpm for 10 min, and the supernatant was kept.  
145 To investigate the possible presence of fatty acid acyl esters, alkaline hydrolysis of the extracts  
146 was performed the same day of analysis by adding 125  $\mu$ L of 2.5 M NaOH in 1.25 mL of extract

147 in a HPLC vial (the same ratio was maintained when hydrolyzing extracts coming from 1 g of  
148 cyclodextrin or resin), vortexing for 0.5 min, and heating at 76 °C for 40 min. Samples were then  
149 cooled at room temperature, neutralized with 125 µL of 2.5 M HCl, and vortexed for 0.5 min.

150 All extracts were passed through 0.2-µm PTFE syringe filters and stored at -20 °C until LC-MS/MS  
151 analysis.

## 152 **2.4. LC-MS/MS analysis**

153 LC-MS/MS analyses were conducted on a 1200 LC system (Agilent Technologies, Santa Clara, CA)  
154 coupled with a 3200 QTRAP triple quadrupole mass spectrometer through a TurboV electrospray  
155 ion source (Applied Biosystems, Foster City, CA), using a previously described methodology  
156 (García-Altarets et al., 2016; Leonardo et al., 2018). Samples were analyzed on an XBridge BEH C8  
157 column, 2.5 µm, 2.1 × 50 mm and an XBridge BEH C8 Prep Guard cartridge, 2.5 µm, 2.1 x 5 mm  
158 (Waters, Milford, MA, USA). A binary gradient was programmed with ultrapure milli-Q water  
159 (mobile phase A) and 90:10 v:v acetonitrile:water (mobile phase B), both containing 6.7 mM of  
160 ammonium hydroxide. Mobile phases were filtered through 0.2-µm nylon membrane filters  
161 (Whatman, Springfield Mill, UK). Chromatographic separations were performed at 30 °C using a  
162 flow rate of 500 µL min<sup>-1</sup>. The elution gradient started at 20% B, reached 100% B in 8 min, held  
163 for 1 min, then back to 20% B in 1 min and equilibrated for 2 min before the next run started.  
164 The injection volume was 10 µL and the auto-sampler was set at 4 °C. A total run time of 12 min  
165 was used. Lipophilic toxins were analyzed in both negative (-ESI) and positive (+ESI) mode,  
166 selecting two product ions per toxin to allow quantification (the most intense transition) and  
167 confirmation (the second intense transitions). Identification was supported by toxin retention  
168 time and multiple reaction monitoring (MRM) ion ratios. Monitored transitions of the detected  
169 toxins were 803.5>255.0 *m/z* (MRM1) and 803.5>113.0 *m/z* (MRM2) for OA, 817.5>255.2 *m/z*  
170 (MRM1) and 817.5>113.1 *m/z* (MRM2) for DTX1, and 876.5>213.3 *m/z* (MRM1) and  
171 876.5>823.5 *m/z* (MRM2) for PTX2. Calibration curves were performed in the range of 2 ng mL<sup>-1</sup>

172 – 40 ng mL<sup>-1</sup> for OA and DTX1, and 5 ng mL<sup>-1</sup> – 50 ng mL<sup>-1</sup> for PTX2, at six calibration levels.  
173 Calibration curve linearities were confirmed before and after each sample set. Curve correlation  
174 coefficients ( $r^2$ ) had to exceed 0.98 and slope deviations had to be below 25% to pursue toxin  
175 quantifications. Limits of detection (LODs, signal/noise > 3) were 1.3 ng/mL for OA and DTX1,  
176 and 1.7 ng/mL for PTX2. Limits of quantification (LOQs, signal/noise > 10) were 4 ng/mL for OA  
177 and DTX1, and 5 ng/mL for PTX2. Samples were analyzed in duplicate.

## 178 **2.5. Statistical analysis**

179 Differences in toxin concentration (*i.e.*, OA and PTX2) among passive sampling disks (*i.e.*,  $\beta$ -CD-  
180 HDI,  $\beta$ -CD-EPI,  $\gamma$ -CD-HDI,  $\gamma$ -CD-EPI and Diaion® HP-20), sampling points, and between weeks  
181 were analyzed with analysis of variance (three-way ANOVA). In addition to  $P$  values, partial eta  
182 squared ( $\eta_p^2$ ) was used as a measure of effect size (*i.e.*, the importance of factors). Similar to  $r^2$ ,  
183 partial  $\eta_p^2$  is the proportion of variation explained for a certain effect, and does not depend on  
184 the number of sources of variation used in the ANOVA, thus it could be compared among  
185 different designs (Tabachnick and Fidell, 2001). In contrast to  $P$  value,  $\eta_p^2$  has the advantage of  
186 allowing the proper comparison of treatments, whereas a lower  $P$  value does not necessarily  
187 mean that a factor has stronger effect (see *e.g.*, Alcaraz et al., 2008). Adjusted (or marginal)  
188 means of a dependent variables are the means for each level of the factor adjusted for the other  
189 variables (see *e.g.*, Alcaraz et al., 2015), and were used to describe differences. Student's  $t$  test  
190 was used to analyze differences in toxin concentration (OA vs. PTX2) for each passive sampling  
191 disk. Quantitative variables were log-transformed prior to analysis because homoscedasticity  
192 and linearity were clearly improved. All statistical analyses were performed with SPSS 26.0.

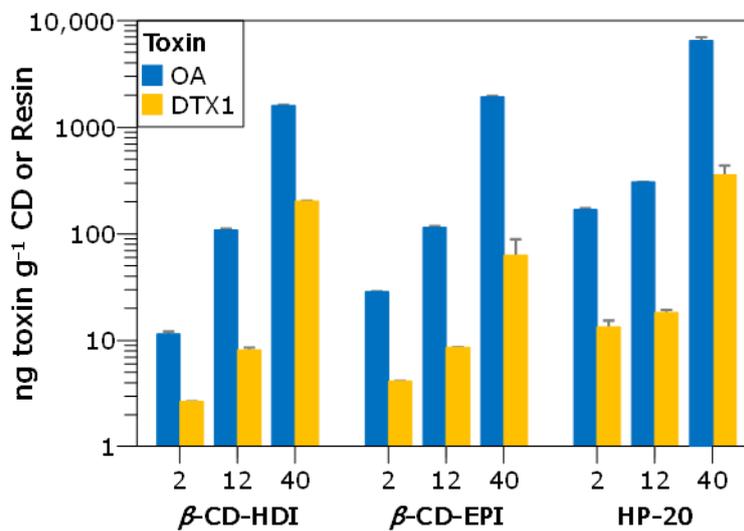
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## 195 **3. Results**

196 **3.1. Analysis of cyclodextrins immersed in *Prorocentrum lima* cultures**

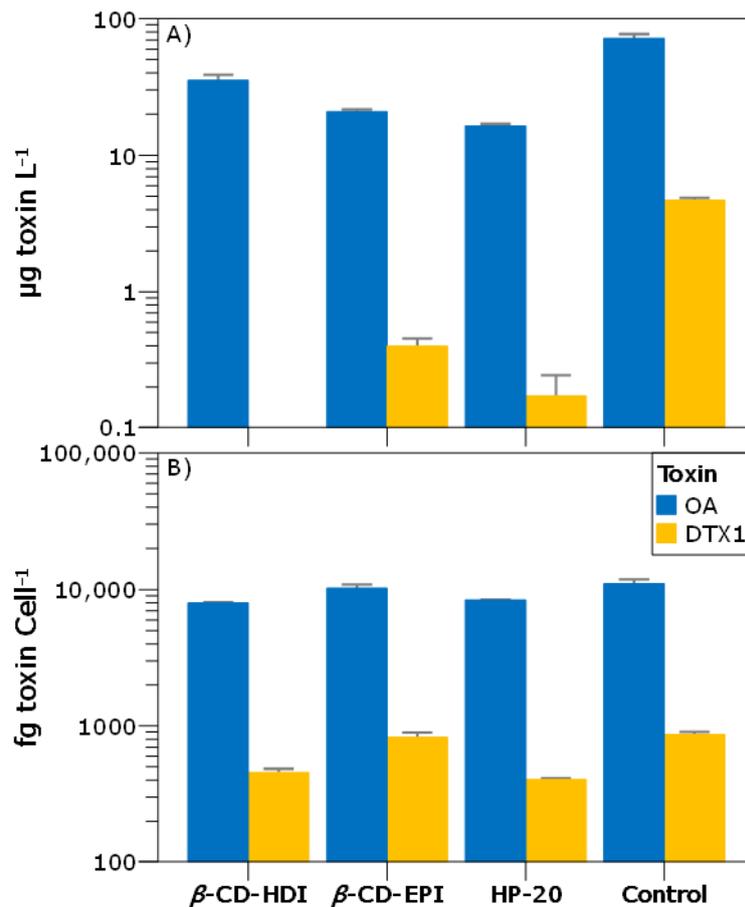
197 Toxins captured by  $\beta$ -CD-HDI,  $\beta$ -CD-EPI and Diaion® HP-20 immersed in *P. lima* cultures were  
198 analyzed by LC-MS/MS, which revealed the presence of free OA and DTX1 in all of them (Fig. 2).  
199 Although it is evident that Diaion® HP-20 provided the highest values, both cyclodextrins showed  
200 OA and DTX1 capture. The OA contents at the different days were 7, 35 and 24% for  $\beta$ -CD-HDI  
201 and 17, 38 and 30% for  $\beta$ -CD-EPI, compared to Diaion® HP-20. The DTX1 contents were 20, 44  
202 and 56% for  $\beta$ -CD-HDI and 31, 47 and 17% for  $\beta$ -CD-EPI, compared to Diaion® HP-20. OA and  
203 DTX1 contents showed exponential trends with culture day.



204

205 **Figure 2.** Free OA and DTX1 captured by the passive sampling disks immersed in *Prorocentrum lima*  
206 cultures and collected at day 2, day 12 and day 40. Each bar corresponds to 1 disk analyzed twice.  
207 *Prorocentrum lima* culture media and the corresponding microalgal cells at the moment of  
208 harvesting were also analyzed (Fig. 3). The control culture, with neither cyclodextrins nor resin,  
209 showed the highest toxin levels in the culture media. Nevertheless, it is interesting to mention  
210 that the total number of cells at harvesting in the control culture was higher than in the cultures  
211 with passive sampling disks, where the number of cells reached 54-65% that of the control  
212 culture (Fig. 3S and Table 1S). The culture media with Diaion® HP-20 showed the lowest OA  
213 contents (23% that of the control), followed by  $\beta$ -CD-EPI (29%) and finally  $\beta$ -CD-HDI (50%) (Fig.  
214 3A), trend that was the opposite of that observed in the passive sampling materials (Fig. 2). Even

215 normalizing to the number of microalgal cells, the trend was the same. DTX1 contents were also  
 216 lower in the culture media with Diaion® HP-20 than in the culture media with  $\beta$ -CD-EPI. However,  
 217 this difference and the lack of DTX1 in the culture media with  $\beta$ -CD-HDI may be simply due to  
 218 the fact that all DTX1 concentrations were very close to the LOD. When observing the toxin  
 219 contents in the microalgal cells, no clear trends were observed (Fig. 3B).



220

221 **Figure 3.** Free OA and DTX1 present in the *P. lima* culture media (A) and in the microalgal cells (B) at day  
 222 40.

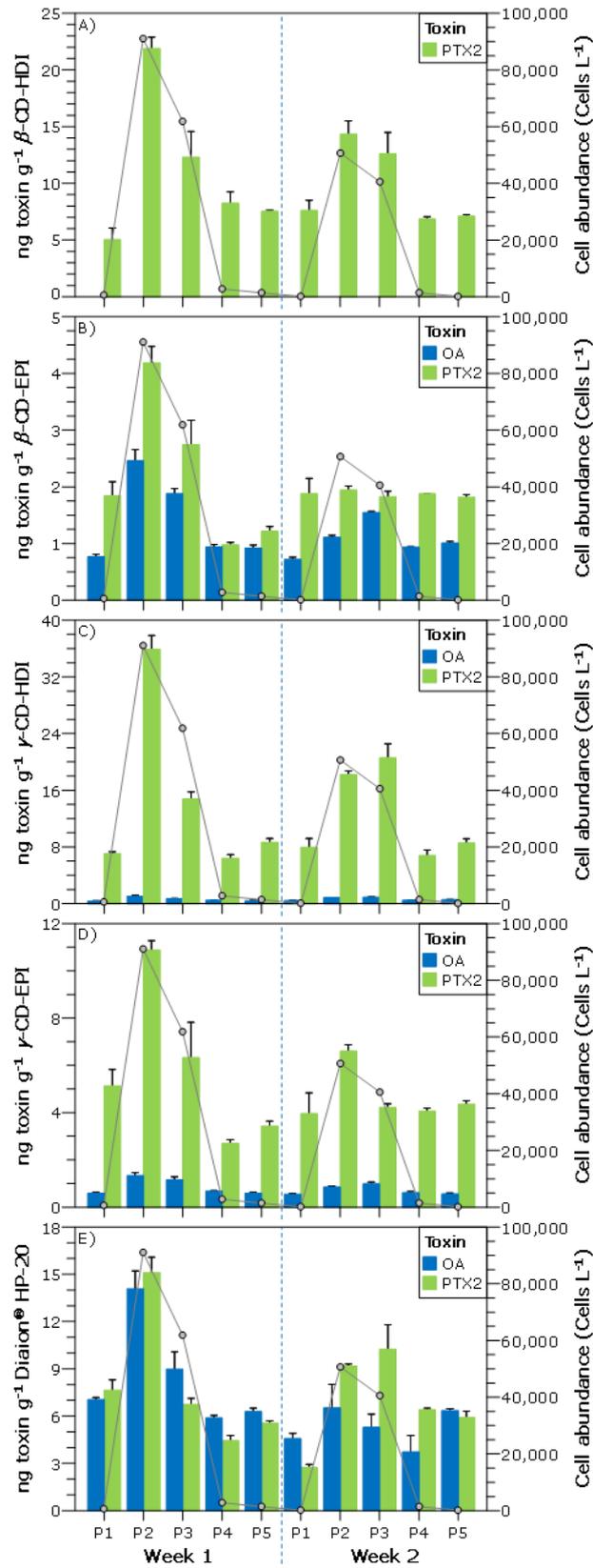
223 The analysis of hydrolyzed extracts showed fatty acid acyl OA esters in all cell pellets and small  
 224 amounts of fatty acid acyl DTX1 esters in the cell pellets corresponding to the culture with  $\beta$ -CD-  
 225 EPI (Fig. 4S). No esters were found in the culture media. Regarding the passive sampling disks,  
 226 OA and DTX1 esters were detected in  $\beta$ -CD-EPI even after only 2 days, and OA esters were  
 227 present in all cyclodextrins and the resin at the last sampling.

228 **3.2. Analysis of cyclodextrins deployed in a harbor during a *Dinophysis sacculus* bloom**

229 Toxins captured by  $\beta$ -CD-HDI,  $\beta$ -CD-EPI,  $\gamma$ -CD-HDI,  $\gamma$ -CD-EPI and Diaion® HP-20 deployed at 5  
230 sampling points of El Masnou harbor (NW Mediterranean Sea) during two consecutive weeks of  
231 a *D. sacculus* bloom were analyzed by LC-MS/MS. The analysis revealed the presence of OA and  
232 PTX2 in all of them, sometimes at very different levels (Fig. 4, Fig. 5 and Fig. 6). Unlike the  
233 experiment in *P. lima* cultures, DTX1 and esters were not found. It is important to mention that  
234 although the LC-MS/MS analysis of extracts from  $\beta$ -CD-HDI revealed presence of OA, the  
235 chromatographic peaks did not fulfil the analytical standard criteria and thus, quantification was  
236 not possible. This effect was probably due to the presence of matrix compounds that interfere  
237 in the analysis for this specific sampling material. Further work would be necessary to remove  
238 this interference and quantify OA.

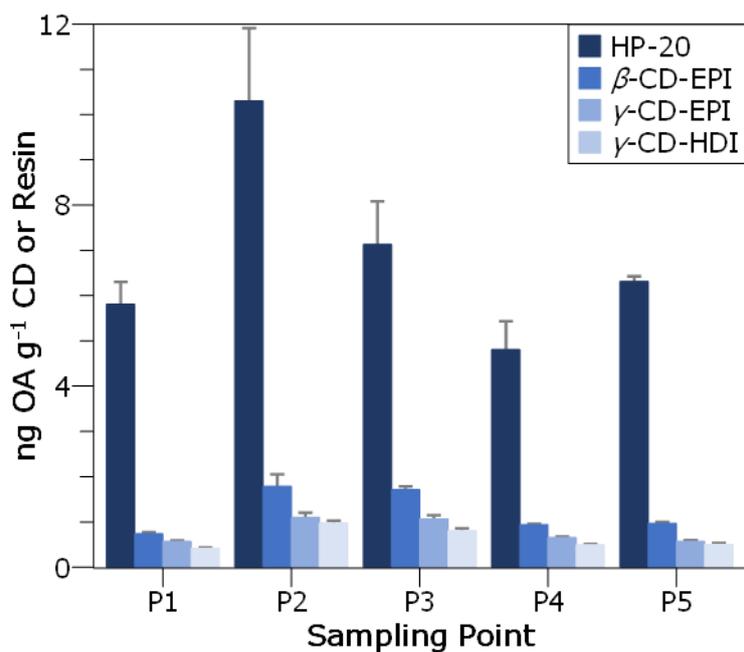
239 In general, considerable differences were observed between sampling points (ANOVA,  $P < 0.0001$   
240 for OA, and  $P < 0.0001$  for PTX2, see Table 2S), P2 showing the highest toxin contents, followed  
241 by P3, and with P1, P4 and P5 showing much lower toxin contents (Table 2S, Fig. 5S and 6S). This  
242 trend is in accordance with the *D. sacculus* cell abundance distribution, which in the first  
243 sampling week was 91,341 cells L<sup>-1</sup> in P2, 62,195 cells L<sup>-1</sup> in P3 and between 880 and 3,040 cells L<sup>-1</sup>  
244 in P1, P4 and P5. Additionally, in P2 and P3, toxin contents in sampling week 1 were higher than  
245 in sampling week 2 (Table 2S, Fig. 7S and 8S), also following the temporal variation of *D. sacculus*  
246 cell abundance, which decreased at the second deployment (e.g. to 50,949 cells L<sup>-1</sup> in P2 and  
247 40,851 cells L<sup>-1</sup> in P3).

248 The global effect of type of cyclodextrin can be observed in Fig. 7, where data points have been  
249 merged. In general terms, the trend for OA contents was Diaion >  $\beta$ -CD-EPI >  $\gamma$ -CD-EPI >  $\gamma$ -CD-  
250 HDI and the trend for PTX2 contents was  $\gamma$ -CD-HDI >  $\beta$ -CD-HDI > Diaion >  $\gamma$ -CD-EPI >  $\beta$ -CD-EPI.  
251 The OA contents were 17, 12 and 9% for  $\beta$ -CD-EPI,  $\gamma$ -CD-EPI and  $\gamma$ -CD-HDI, respectively,  
252 compared to Diaion® HP-20. The PTX2 contents were 182, 144, 70 and 27% for  $\gamma$ -CD-HDI,  $\beta$ -CD-  
253 HDI,  $\gamma$ -CD-EPI and  $\beta$ -CD-EPI, respectively, compared to Diaion® HP-20. Therefore, although  
254 Diaion® HP-20 is better to capture OA, CD-HDIs are more efficient for PTX2.



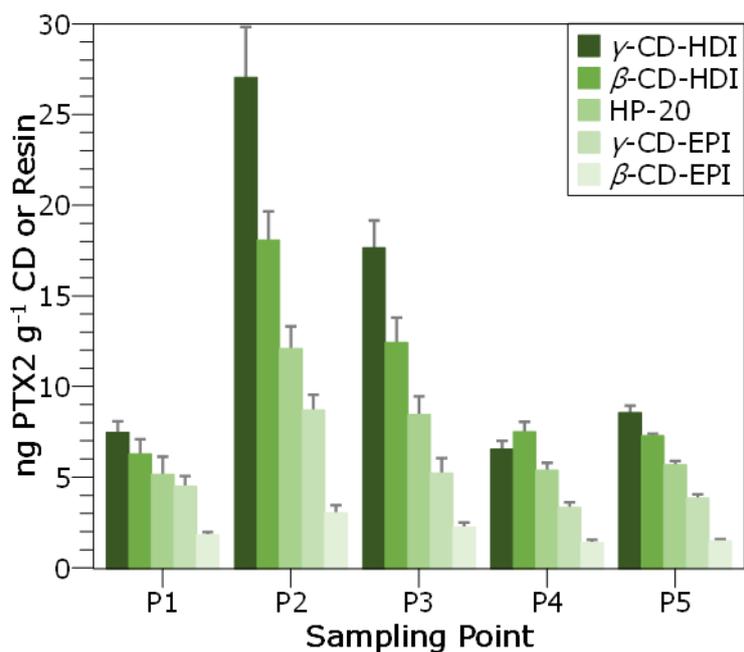
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256 **Figure 4.** Left axis: OA and PTX2 captured by the passive sampling disks deployed El Masnou harbor (NW  
 257 Mediterranean Sea). Right axis: *Dinophysis sacculus* cell abundance in the water column (average from  
 258 deployment and collection days). Each bar corresponds to the mean of 2 disks (duplicates) analyzed  
 259 twice.



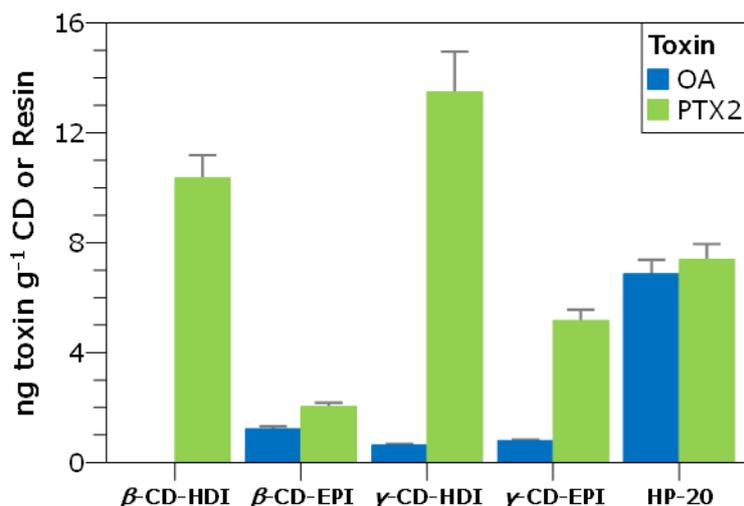
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261 **Figure 5.** OA captured by the passive sampling disks per cyclodextrin/resin and sampling point (P1 to P5)  
 262 of El Masnou harbor (NW Mediterranean Sea). Each bar corresponds to the mean of 4 disks (duplicates  
 263 and 2 weeks) analyzed twice.



264

265 **Figure 6.** PTX2 captured by the passive sampling disks per cyclodextrin/resin and sampling point (P1 to  
 266 P5) of El Masnou harbor (NW Mediterranean Sea). Each bar corresponds to the mean of 4 disks  
 267 (duplicates and 2 weeks) analyzed twice.



268

269 **Figure 7.** Total OA and PTX2 captured by the passive sampling disks per cyclodextrin/resin deployed at El  
 270 Masnou harbor (NW Mediterranean Sea). Each bar corresponds to the mean of 20 disks (duplicates, 2  
 271 weeks and 5 points) analyzed twice. Error bars are the standard error of the means.

272

## 273 4. Discussion

### 274 4.1. Cyclodextrins immersed in *Prorocentrum lima* cultures

275 The immersion of β-CD-HDI, β-CD-EPI and Diaion® HP-20 in *P. lima* cultures for different days  
 276 provided information about the toxin production of the strain combined with the toxin capture  
 277 efficiency of the different passive sampling materials. LC-MS/MS analysis revealed the presence  
 278 of OA and DTX1 in all extracts even at day 2 and, as expected, toxin contents showed an  
 279 exponential increase trend with culture day. Although Diaion® HP-20 provided the highest toxin  
 280 contents, both cyclodextrins also showed OA and DTX1 capture. Except for DTX1 at day 40, β-  
 281 CD-EPI was slightly more efficient in capturing toxins than β-CD-HDI, and this observation was  
 282 more evident at day 2 than at days 12 and 40, suggesting a faster capture rate of the former  
 283 cyclodextrin. The percentage decrease of DTX1 in β-CD-EPI at day 40 (17%) could be due to  
 284 different experimental parameters, such as a lower DTX1 production at that stage of the culture  
 285 or the saturation of the passive sampling material. In general terms, although DTX1 values were  
 286 around 10-fold lower than OA values, DTX1 capture with cyclodextrins was more efficient than  
 287 OA capture compared to Diaion® HP-20. The additional methyl moiety in one of the extremes of

288 DTX1 (Fig. 9S) makes it less polar than OA, characteristic that could be favoring the interaction  
289 with the cyclodextrins, as previously observed for Diaion® HP-20 (Li et al., 2011).

290 When making all these comparisons, it is important to keep in mind that  $\beta$ -CD-HDI,  $\beta$ -CD-EPI and  
291 Diaion® HP-20 were immersed in different glass bottles during culture, in order to avoid  
292 competition between them. Therefore, differences in microalgae growth and toxin composition  
293 of the media cannot be discarded. Analysis of the culture media and the corresponding  
294 microalgal cells contributed to better characterize the system. The analysis of *P. lima* culture  
295 media at the moment of harvesting revealed lower dissolved toxin contents in the cultures with  
296 passive sampling disks than in the control. However, the number of *P. lima* cells was also lower.  
297 A possible explanation is that the presence of the disks may have inhibited the culture growth  
298 (Table 1S), by decreasing light exposure or capturing culture media components (e.g. nutrients,  
299 vitamins and metals). For those cultures with passive sampling disks, the trend of dissolved OA  
300 content in the media was inversely proportional to that in the passive sampling material,  
301 observation that supports the capture efficiency previously observed, which is Diaion® HP-20 >  
302  $\beta$ -CD-EPI >  $\beta$ -CD-HDI. Regarding the toxin contents in microalgal cells, no clear trends were  
303 observed among passive sampling materials and no significant differences with the control. This  
304 seems to indicate that the passive sampling disks, although are affecting culture growth, may  
305 not be affecting toxin production per cell.

306 Regarding to absolute values and compared to other works, much lower toxin contents are  
307 obtained herein (e.g. 168 ng OA g<sup>-1</sup> and 13 ng DTX1 g<sup>-1</sup> for the Diaion® HP-20 after 48 h in front  
308 of 982 ng OA g<sup>-1</sup> and 846 ng DTX1 g<sup>-1</sup> found by Fux et al., 2008), differences probably due to the  
309 strain, its age and the culture experimental conditions.

310 *Prorocentrum lima* is a well-known producer of OA, DTX1, DTX2 and several types of esters, the  
311 toxin profile and contents depending on the strain and growing conditions (Pan et al., 1999,  
312 Bravo et al., 2001, Nascimento et al., 2005, Paz et al., 2007, Morton et al., 2009, Vale et al., 2009,

313 Li et al., 2012, Hu et al., 2017). However, no esters or only trace amounts are usually found in  
314 culture media (Pan et al., 1999, Nascimento et al., 2005), probably because they are hydrolyzed  
315 before their release from the cells (Hu et al., 2017). In the current work, esters were not found  
316 in the culture media either. Regarding their presence in passive samplers, a previous work  
317 showed that OA esters were detected in SPATT bags with Diaion® HP-20 (Mackenzie et al., 2011).  
318 In the current work, OA and DTX1 esters were detected in some passive sampling disks, mainly  
319 in those from the last sampling. The lipophilic character of the fatty acid acyl ester derivatives  
320 makes them suitable to be captured by the cyclodextrins and the Diaion® HP-20 resin. Therefore,  
321 these passive sampling materials could be pre-concentrating esters once released from the cells,  
322 allowing their detection better than in the media.

#### 323 **4.2 Cyclodextrins deployed in a harbor during a *Dinophysis sacculus* bloom**

324 The deployment of  $\beta$ -CD-HDI,  $\beta$ -CD-EPI,  $\gamma$ -CD-HDI,  $\gamma$ -CD-EPI and Diaion® HP-20 passive samplers  
325 during a *D. sacculus* bloom revealed the capture of OA and PTX2. However, there is not a clear  
326 indication about which was the most predominant toxin, since different passive sampling  
327 materials provide different trends (Fig. 4). Additionally, in the current work, toxin levels in the  
328 cyclodextrins (or in the commercial resin used as a control) did not reach the values found in  
329 other works, which may suggest an overall lower toxin production from the bloom. The toxin  
330 concentration and profile of *D. sacculus* may vary depending on the strain, its geographical  
331 origin, the experimental parameters of the culture and, of course, multiple environmental  
332 conditions in the case of field samples. In the Mediterranean, *D. sacculus* blooms have been  
333 associated to OA (Cañete et al., 2008; Garibo et al., 2014; García-Altare et al., 2016; Leonardo  
334 et al., 2018; Giacobbe et al., 2000), PTX2 (Cañete et al., 2008; García-Altare et al., 2016; and  
335 sometimes DTX1, to a lesser extent (Giacobbe et al., 2000). In a previous study performed by our  
336 group during a *D. sacculus* bloom in Alfacs Bay (NW Mediterranean Sea), 200 km to the south of  
337 Masnou harbor, PTX2 was the main component in the toxin profiles of phytoplankton aggregates  
338 (up to 668 pg PTX2 cell<sup>-1</sup> in front of 461 pg OA cell<sup>-1</sup>), while OA was the most concentrated toxin

339 in Diaion® HP-20 SPATTs (94 ng OA g<sup>-1</sup> in front of 42 ng PTX2 g<sup>-1</sup>) (García-Altare et al., 2016). In  
340 another work performed with a culture of a *D. sacculus* strain from Galicia (NE Atlantic Ocean),  
341 PTX2 was also the main toxin in the cells (13 pg cell<sup>-1</sup>), followed by OA (8 pg cell<sup>-1</sup>), whereas OA  
342 was more abundant in the medium (28 ng OA mL<sup>-1</sup> in front of 23 ng PTX2 mL<sup>-1</sup>) (Riobó et al.,  
343 2013). In that work, only traces of DTX1 were observed in the cell pellet extract, but not in the  
344 medium. OA contents in SPATT discs deployed during a *D. acuta* event in the west coast of Ireland  
345 (Atlantic Ocean) were also much higher than PTX2, followed by DTX2 (maximum of 5645 ng OA g<sup>-1</sup>,  
346 1265 ng PTX2 g<sup>-1</sup> and 533 ng DTX2 g<sup>-1</sup>) (Fux et al., 2009). The deployment of passive samplers  
347 in Galicia (NE Atlantic Ocean) during a bloom where several *Dinophysis* species were present  
348 (mainly *D. ovum* and *D. acuminata*, but also *D. acuta* and *D. caudata*, depending on the day) also  
349 followed this pattern (maximum 4495 ng OA g<sup>-1</sup>, 2705 ng PTX2 g<sup>-1</sup> and 1876 ng DTX2 g<sup>-1</sup>) (Pizarro  
350 et al., 2013). On the contrary, toxin contents found in SPATT discs deployed during another  
351 *D. acuta* bloom near the south-west coast of Ireland (Atlantic Ocean) were higher for PTX2  
352 (between 1.1 and 4.5 µg g<sup>-1</sup>), followed by DTX2 (between 0.9 and 3.7 µg g<sup>-1</sup>) and finally OA  
353 (between 0.8 and 2.9 µg g<sup>-1</sup>) (Fux et al., 2010), and the deployment of SPATT bags during a  
354 *D. acuminata* bloom in New Zealand also resulted in higher levels of PTX2 compared to OA and  
355 DTX1, which were 1.73 µg PTX2 g<sup>-1</sup> and 0.11 µg OA/DTX1 g<sup>-1</sup> when the maximum cell numbers of  
356 *D. acuminata* were observed. All these works underline the complexity in the understanding and  
357 elucidation of toxin profiles.

358 Nevertheless, results provided herein showed interesting and informative trends. Regarding  
359 location distribution, the highest toxin contents, for both OA and PTX2, were detected in the  
360 sampling points with the highest *D. sacculus* cell abundance (Fig. 4). When looking at the  
361 geographical position of the sampling points (Fig. 2S), P2 is in an interior corner of the harbor,  
362 followed by P3, and finally by P4, P5 and P1. Thus, a correlation with the spatial *D. sacculus*  
363 bloom dynamics is observed, *D. sacculus* cells accumulating in the most confined part of the  
364 harbor. Additionally, apart from the *D. sacculus* cell abundance, the diffusion of the dissolved

365 toxins may be playing a role. Regarding temporal distribution, data from P2 and P3 clearly  
366 demonstrate that toxin contents correlate with the temporal variation of *D. sacculus* cell  
367 abundance. Another important issue is that OA and PTX2 were detected in almost all samples,  
368 even at very low *D. sacculus* cell abundances (as low as 400 cells L<sup>-1</sup>), which indicated how  
369 common traces of these toxins are in seawater and how stable they are (already demonstrated  
370 at least for OA, Blanco et al., 2018), and also how sensitive the SPATT strategy is.

371 It is evident that PTX2 capture with cyclodextrins was more efficient than OA capture (Fig. 7).  
372 This can be explained considering the structures of both toxins and cyclodextrin polymers and  
373 the interactions involved in the capture process. The Diaion® HP-20 resin is known to capture a  
374 wide range of organic molecules from aqueous solution through a hydrophobic binding  
375 mechanism involving its styrene-divinylbenzene matrix. That is why it showed no significantly  
376 different OA and PTX2 contents (Student's *t*-test, *P* = 0.53) and is not selective to any of the toxins  
377 as compared with the cyclodextrin-based materials that show higher contents for PTX2  
378 (Student's *t*-test, *P* < 0.0001 in all cases). Unlike OA and DTX1, OA and PTX2 have very different  
379 structures with different overall hydrophilic/hydrophobic balances (Fig. 9S). PTX2 is a neutral  
380 molecule with a ring structure composed by oxolane and oxane rings connected by aliphatic  
381 spacers (O'Rourke et al., 2017). This makes PTX2 a polar but essentially hydrophobic molecule,  
382 which could favor its capture by cyclodextrins and may explain the higher capture efficiency of  
383 the cyclodextrin polymers over Diaion® HP-20. On the other hand, OA is also composed by  
384 oxolane and oxane rings but in a linear structure and possesses an ionizable carboxylic acid group  
385 at the A ring resulting in a higher polarity and solubility in aqueous solution (Mackenzie et al.,  
386 2004). This may explain, in general terms, its lower capture by the cyclodextrins as compared to  
387 the commercial resin.

388 In the case of the cyclodextrin materials, their capture efficiency should be a combination of the  
389 cavity size and the nature of the bridging units. Regarding OA and cavity size, capture efficiencies  
390 with  $\beta$ -CD-EPI, consisting of seven glucopyranose units, were only slightly higher (Table 2S and

391 Fig. 10S) than with  $\gamma$ -CD-EPI, consisting of eight glucopyranose units (it is important to mention  
392 that absolute OA contents were in general very low). Therefore, the cavity size plays a negligible  
393 effect, if any, in the capture process of OA. Nevertheless, when looking at the PTX2 results,  $\gamma$ -CD-  
394 HDI and  $\gamma$ -CD-EPI provided significantly higher capture efficiencies than their  $\beta$ -CD counterparts  
395 (Table 2S and Fig. 11S), suggesting that the larger internal diameter (0.95 nm for  $\gamma$ -CDs in front  
396 of 0.78 nm for  $\beta$ -CDs (Szejtli, 1998)) could be playing a role in accommodating the PTX2 molecule,  
397 most likely through the D/E rings.

398 On the other hand, the role of the bridging units was more evident in the case of PTX2 than OA.  
399 Both CD-EPIs, which contain polar hydroxyalkyl groups of different lengths depending on the  
400 degree of polymerization, were slightly more efficient than CD-HDI in capturing the more polar  
401 OA (Table 2S and Fig. 10S). On the contrary, CD-HDIs were much more efficient in capturing the  
402 more lipophilic PTX2 than CD-EPIs (Table 2S and Fig. 11S). The explanation could rely on the  
403 higher hydrophobicity of the HDI spacer (which contains six CH<sub>2</sub> groups connected to the CD by  
404 O(C=O)NH groups) compared to EPI that provides a more hydrophobic environment for PTX2, as  
405 well as the occurrence of hydrogen bond interactions between the amido and OH groups of the  
406 capture polymer with the polar groups of the toxin. Therefore, the nature of the bridging units  
407 is certainly playing a more decisive role than the cavity size in this case as evidenced by the trend  
408  $\gamma$ -CD-HDI >  $\beta$ -CD-HDI > Diaion >  $\gamma$ -CD-EPI >  $\beta$ -CD-EPI. Another important point to consider is that,  
409 in contrast to Diaion® HP-20,  $\gamma$ -CD-HDI and  $\beta$ -CD-HDI are not porous materials and the capture  
410 of PTX2 occurs mainly on the surface of the particles. Hence, the high capture efficiency showed  
411 by materials with a lower active surface such as both cyclodextrin polymers is indicative of the  
412 high strength of the intermolecular interactions involved.

413 An experiment was performed in the lab, with seawater spiked with the two toxins at equimolar  
414 concentrations, to investigate if competition of OA and PTX2 for the cyclodextrin cavities could  
415 explain the results obtained in the *in-situ* deployment experiment. However, all cyclodextrins  
416 were able to capture both toxins with recovery values of about 80%, and no competition was

417 observed. In fact, although the saturation of the cyclodextrins was not evaluated, it is evident  
418 that toxin contents in the cyclodextrins from the *in-situ* deployment experiment were far from  
419 saturation (much lower toxin contents than those reached in the culture experiment). Thus, what  
420 is happening in nature is much more complex and other environmental parameters are certainly  
421 playing a significant role. Since cyclodextrins are not specific for marine toxins, capture and  
422 competition of other compounds from seawater (e.g. chemical contaminants, organic matter,  
423 micro/nanoplastics and demucilaged seeds) cannot be discarded.

#### 424 **4.3 General discussion**

425 The cyclodextrin-based materials tested in this work have provided useful information regarding  
426 the toxin profile of a *P. lima* strain and the spatial and temporal dynamics of a *D. sacculus* bloom.  
427 It is necessary to take into account that these passive sampling materials are capturing dissolved  
428 toxins from the culture media or the seawater, during a time interval, and with different  
429 efficiencies. Therefore, as observed in this work, discrepancies may arise among them. Although  
430 more work is required to better understand the results and to fully characterize the two case  
431 studies, multiple experimental parameters are not under control (e.g. diffusion of dissolved  
432 toxins, competition of toxins with other compounds that can also be captured by cyclodextrins,  
433 stability of dissolved toxins, environmental conditions, etc.). Nevertheless, this work is a step  
434 forward to understand what is happening in nature.

435 Compared to other passive sampling materials, cyclodextrins have the advantages of being easy  
436 to manufacture, cheap and ecologically sustainable, and their chemical and structural versatility  
437 could allow a rational synthesis according to the type of toxin to be captured. Therefore, they  
438 could be useful as early warning tools in monitoring programs. To this purpose, toxin contents in  
439 shellfish from specific regions should be determined, tailor-made cyclodextrins should be tested,  
440 and correlation between both should be established.

441

## 442 **5. Conclusions**

443 Summarizing, the results of the experiments described herein demonstrate the potential of  
444 cyclodextrin polymers as new materials for passive sampling of lipophilic marine toxins. In the  
445 culture experiment, OA and DTX1, and related esters at a lower extent, have been captured along  
446 *P. lima* culture time, the first signals being detected even after only 2 days. In the *in-situ*  
447 deployment experiment, toxin contents were significantly higher in the sampling points and  
448 dates where *D. sacculus* cell abundance was also higher, the effect being more evident for PTX2  
449 than for OA. While the exact capture mechanism and the role of cavities and spacers is currently  
450 under study, the evaluated cyclodextrin-based materials have already proven efficient in  
451 providing integrated contents of toxins released into culture media and the environment. Further  
452 investigation is underway to evaluate the capture and equilibrium rates, saturation, regeneration  
453 and reusability, to remove matrix effects and to establish the correlations with shellfish  
454 contamination with the aim to develop novel passive sampling materials able to satisfy the  
455 environmental monitoring demands of specific geographic regions.

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## 602 **Acknowledgments**

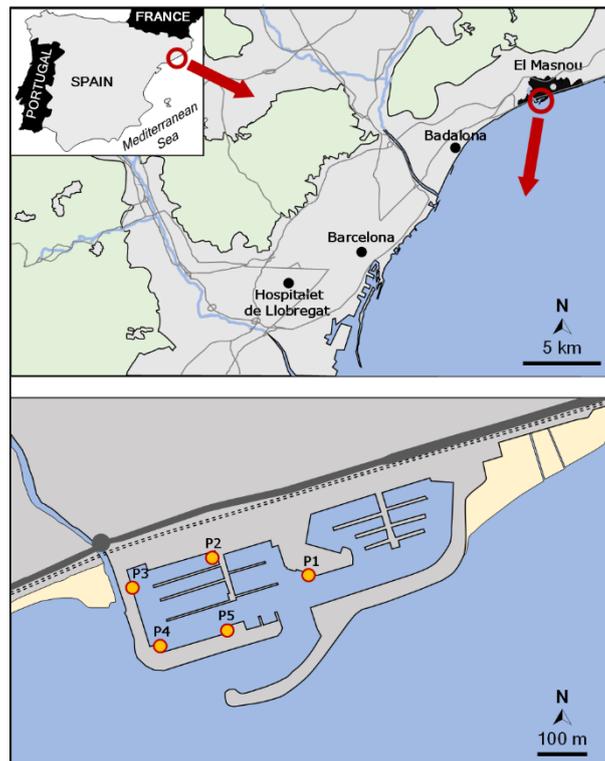
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## Supplementary information

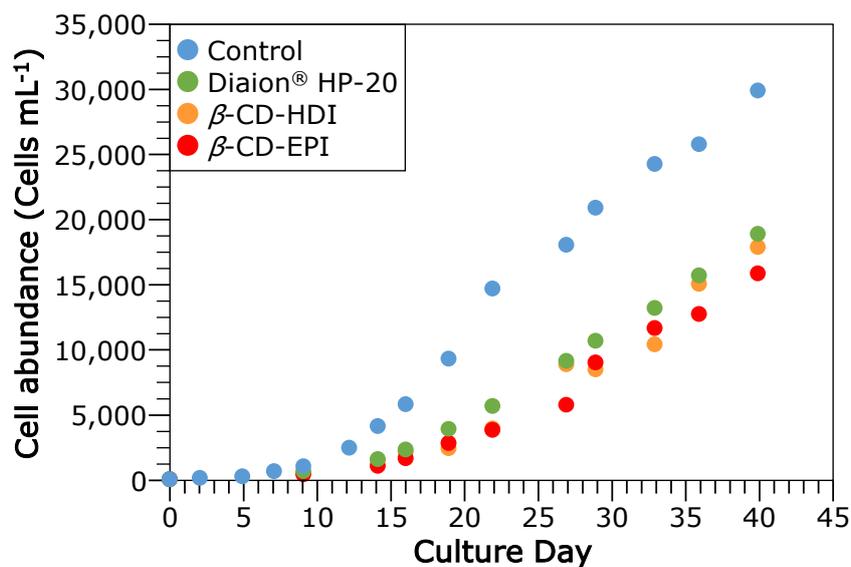


**Figure 1S.** Passive sampling disks containing  $\beta$ -CD-HDI,  $\beta$ -CD-EPI,  $\gamma$ -CD-HDI,  $\gamma$ -CD-EPI or Diaion® HP-20 (A), immersed in *P. lima* cultures (B), and deployed in El Masnou harbor waters (C).



**Figure 2S.** Map of the Iberian Peninsula with sampling points at El Masnou harbor (41.47614 N, 2.32244 E).

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**Figure 3S.** Growth curves for the *P. lima* cultures of strain IRTA-SMM-17-47 in the presence of  $\beta$ -CD-HDI,  $\beta$ -CD-EPI and Diaion® HP-20 as well as for the control.

**Table 1S.** Generalised von Bertalanffy growth-model parameter estimates for the *P. lima* cultures of strain IRTA-SMM-17-47. Model parameters were fitted by non-linear regression using the Levenberg-Marquardt algorithm.

CD or Resin	Akaike information criterion	Growth rate		Carrying capacity	
		Estimate	Standard error	Estimate	Standard error
Control	1032.05	0.09703	0.01133	33668.40	2464.46
Diaion® HP-20	665.33	0.06112	0.00699	34117.65	4459.12
$\beta$ -CD-HDI	707.28	0.06785	0.01399	27540.33	6020.69
$\beta$ -CD-EPI	684.41	0.06148	0.01087	31269.90	6821.80

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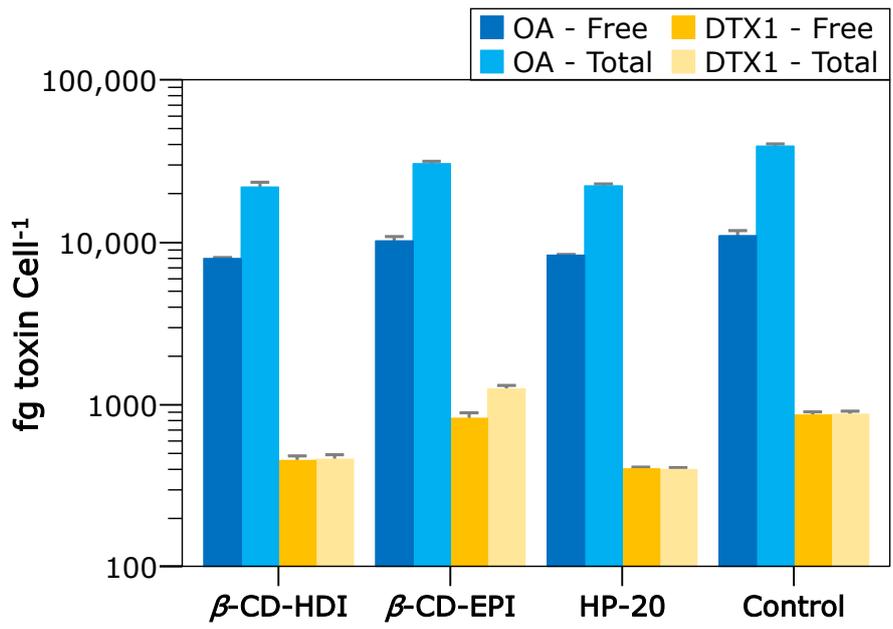


Figure 4S. Free and total OA and DTX1 present in the *P. lima* cells at day 40.

60 **Table 2S.** Three-way ANOVAs on the effects of passive sampling disk, sampling point and week  
 61 on toxin concentration (A), and ANOVAs on the effects of passive sampling disk, zone (Z1 =  
 62 sampling points P2 and P3, Z2 = sampling points P1, P4 and P5) and week on toxin  
 63 concentration (B).

Factor	Toxin							
	OA				PTX2			
	F	df	P	$\eta_p^2$	F	df	P	$\eta_p^2$
<b>A) Among Points</b>								
Disk	1417	3, 120	<0.0001	0.973	502.2	4, 150	<0.0001	0.931
Point	97.54	4, 120	<0.0001	0.765	163.3	4, 150	<0.0001	0.813
Week	40.84	1, 120	<0.0001	0.254	5.395	1, 150	0.022	0.035
Disk × Point	4.243	12, 120	<0.0001	0.298	4.976	16, 150	<0.0001	0.347
Disk × Week	18.15	3, 120	<0.0001	0.312	0.674	4, 150	0.611	0.018
Point × Week	14.48	4, 120	<0.0001	0.326	26.01	4, 150	<0.0001	0.410
Disk × Point × Week	1.337	12, 120	0.207	0.118	1.514	16, 150	0.101	0.139
<b>B) Between Zones</b>								
Disk	894.7	3, 144	<0.0001	0.949	222.96	4, 180	<0.0001	0.832
Point	251.2	1, 144	<0.0001	0.636	250.93	1, 180	<0.0001	0.582
Week	35.58	1, 144	<0.0001	0.198	5.440	1, 180	0.021	0.029
Disk × Point	3.300	3, 144	0.022	0.064	5.135	4, 180	0.001	0.102
Disk × Week	12.82	3, 144	<0.0001	0.211	0.151	4, 180	0.962	0.003
Point × Week	17.26	1, 144	<0.0001	0.107	16.97	1, 180	<0.0001	0.086
Disk × Point × Week	1.505	3, 144	0.216	0.030	1.167	4, 180	0.327	0.025

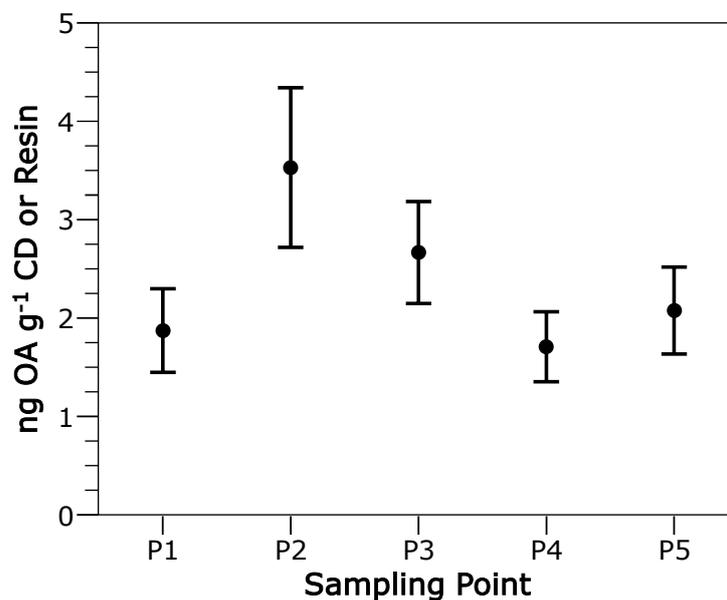
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65 Okadaic acid concentration differed significantly among passive sampling disks, sampling points  
 66 and weeks. Sampling points P2 and P3 showed the highest OA values (Fig. 5S), which decreased  
 67 the second week (Fig. 7S). Among passive sampling disks, Diaion® HP-20 showed the highest  
 68 values, followed (but much lower) by  $\beta$ -CD-EPI,  $\gamma$ -CD-EPI, and  $\gamma$ -CD-HDI (Fig. 10S). This pattern  
 69 was similar in all sampling points (Fig. 5). There was also disk × point, disk × week, and point ×  
 70 week interactions, which can be explained by small deviations from the general pattern of OA  
 71 concentration among passive sampling disks, sampling points and weeks. According to  $\eta_p^2$   
 72 values, differences in OA values were mainly explained by the passive sampling disk, followed  
 73 by sampling point, since week and the significant interactions had a minor weight.

74 Pectenotoxin-2 concentration differed significantly among passive sampling disks, sampling  
 75 points and weeks. Sampling points P2 and P3 showed the highest concentration values (Fig. 6S),  
 76 which decreased the second week (Fig. 8S). Among passive sampling disks,  $\gamma$ -CD-HDI showed the  
 77 highest values, followed by  $\beta$ -CD-HDI, Diaion® HP-20,  $\gamma$ -CD-EPI, and  $\beta$ -CD-EPI (Fig. 11S). This  
 78 pattern was similar in all sampling points (Fig. 6). There was also disk × point and point × week  
 79 interactions, which can be explained by small deviations from the general pattern of PTX2  
 80 concentration among passive sampling disks, sampling points and weeks. According to  $\eta_p^2$   
 81 values, passive sampling disk and sampling point were the most important factors in explaining  
 82 differences in PTX2 values since week and the significant interactions had a minor weight.

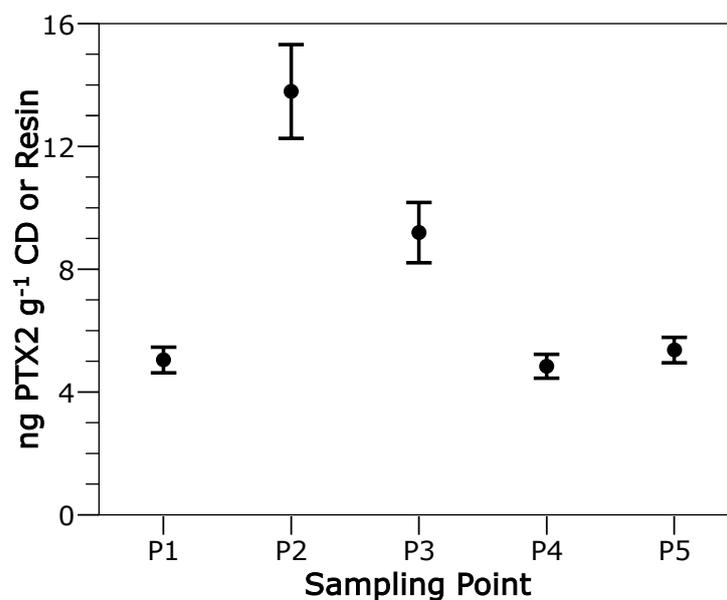
83 When sampling points were pooled in two different zones (zone 1 including sampling points 2  
 84 and 3, and zone 2 including sampling points 1, 4 and 5) according to toxin concentration, the  
 85 same results were obtained.

86 **Figure 5S.** Total OA captured by the passive sampling disks per sampling point (P1 to P5) of El Masnou  
87 harbor (NW Mediterranean Sea). Each point corresponds to the mean of 20 disks (duplicates, 2 weeks  
88 and 4 cyclodextrins/resin) analyzed twice. Error bars are the standard error of the means.



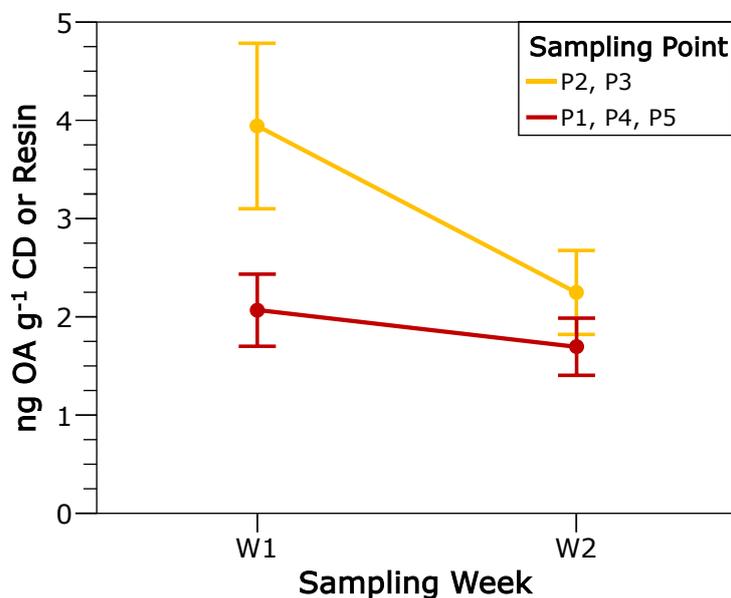
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104 **Figure 6S.** Total PTX2 captured by the passive sampling disks per sampling point (P1 to P5) of El Masnou  
105 harbor (NW Mediterranean Sea). Each point corresponds to the mean of 20 disks (duplicates, 2 weeks  
106 and 5 cyclodextrins/resin) analyzed twice. Error bars are the standard error of the means.

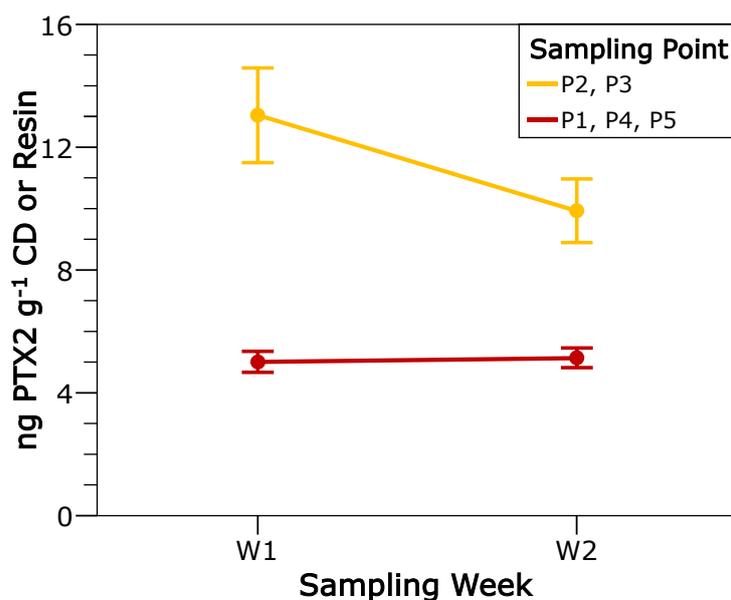


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122 **Figure 7S.** Total OA captured by the passive sampling disks per sampling week (W1 and W2) and  
123 sampling point group of El Masnou harbor (NW Mediterranean Sea). Each point corresponds to the  
124 mean of 25 disks (duplicates, 2 or 3 sampling points and 5 cyclodextrins/resin) analyzed twice. Error bars  
125 are the standard error of the means.

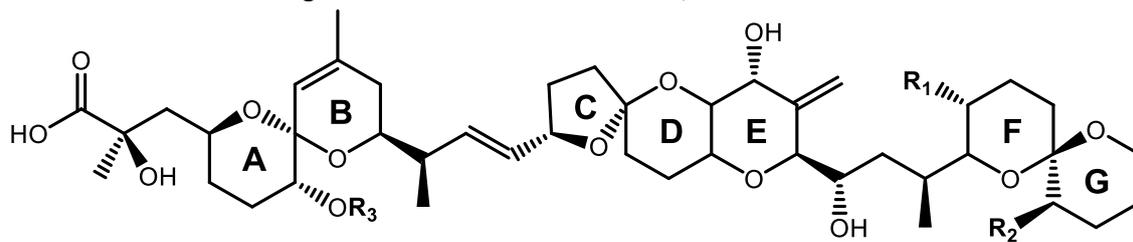


139 **Figure 8S.** Total PTX2 captured by the passive sampling disks per sampling week (W1 and W2) and  
140 sampling point group of El Masnou harbor (NW Mediterranean Sea). Each point corresponds to the  
141 mean of 25 disks (duplicates, 2 or 3 sampling points and 5 cyclodextrins/resin) analyzed twice. Error bars  
142 are the standard error of the means.

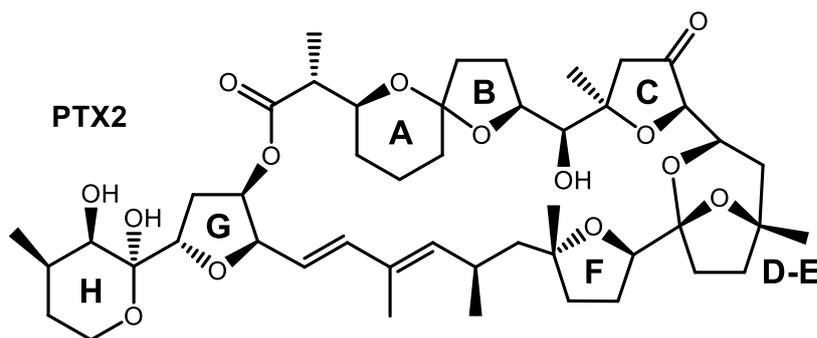


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Figure 9S. Chemical structures of OA, DTX1 and PTX2.



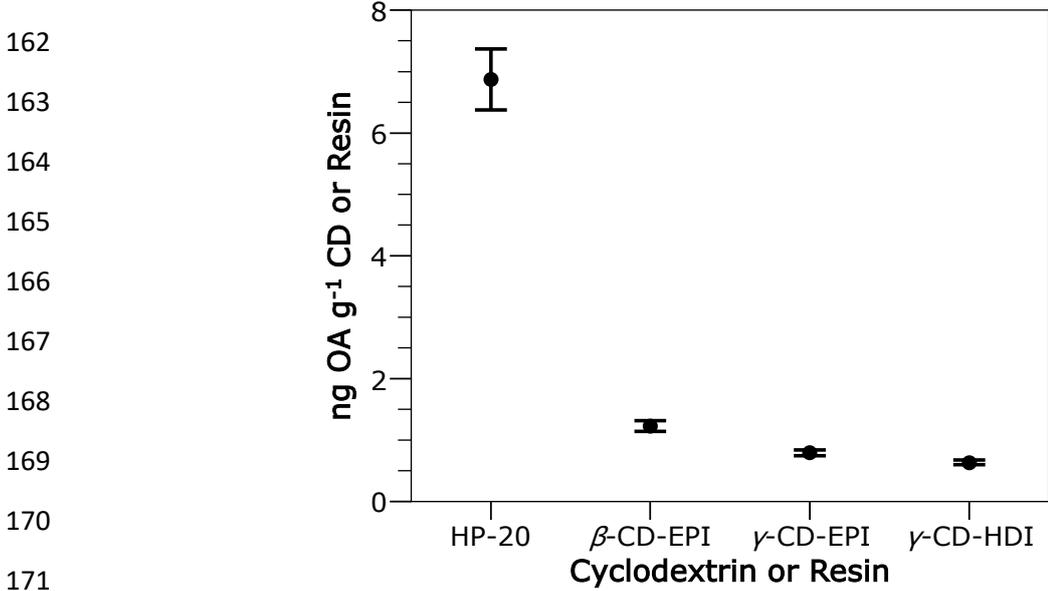
OA: R<sub>1</sub> = CH<sub>3</sub>, R<sub>2</sub> = H, R<sub>3</sub> = H  
DTX1: R<sub>1</sub> = CH<sub>3</sub>, R<sub>2</sub> = CH<sub>3</sub>, R<sub>3</sub> = H



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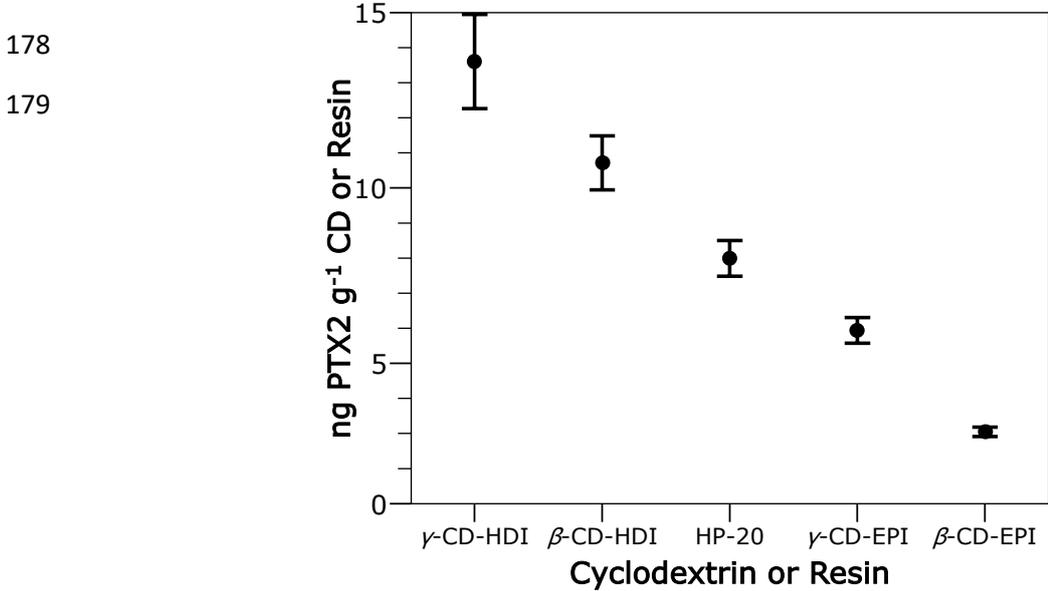
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159 **Figure 10S.** Total OA captured by the passive sampling disks. Each point corresponds to the mean of 40  
160 disks (4 replicates, 2 weeks and 5 sampling points). Error bars are the standard error of the means.  
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175 **Figure 11S.** Total PTX2 captured by the passive sampling disks. Each point corresponds to the mean of  
176 40 disks (4 replicates, 2 weeks and 5 sampling points). Error bars are the standard error of the means.  
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